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1 Defining “serious harm” to the marine environment in the context of deep-seabed mining

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37 Highlights

- 38 • There is growing likelihood of minerals mining in the deep sea. (64)
- 39
- 40 • Assessing the significance of resulting environmental impacts takes on urgency. (79)
- 41
- 42 • The ISA is developing regulations for seabed mining which must prevent serious harm.
- 43
- 44 • Defining “serious harm” is critical to effective regulation of mining activities. (82)
- 45
- 46 • Deep faunal vulnerabilities derive from low growth rates, species longevity and rarity.
- 47
- 48 • Connectivity, resilience, and cumulative impacts are key to significance assessment.
- 49

50 ABSTRACT

51 Increasing interest in deep-seabed mining has raised many questions surrounding its potential  
52 environmental impacts and how to assess the impacts’ significance. Under the United Nations Convention  
53 on the Law of the Sea (UNCLOS), the International Seabed Authority (ISA) is charged with ensuring  
54 effective protection of the marine environment as part of its responsibilities for managing mining in  
55 seabed areas beyond national jurisdiction (the Area) on behalf of humankind. This paper examines the  
56 international legal context for protection of the marine environment and defining the significant adverse  
57 change that can cause “serious harm”, a term used in the ISA Mining Code to indicate a level of harm that  
58 strong actions must be taken to avoid. It examines the thresholds and indicators that can reflect  
59 significant adverse change and considers the specific vulnerability of the four ecosystems associated with  
60 the minerals targeted for mining: (1) manganese (polymetallic) nodules, (2) seafloor massive  
61 (polymetallic) sulphides, (3) cobalt-rich (polymetallic) crusts and (4) phosphorites. The distributions and  
62 ecological setting, probable mining approaches and the potential environmental impacts of mining are  
63 examined for abyssal polymetallic nodule provinces, hydrothermal vents, seamounts and phosphorite-rich  
64 continental margins. Discussion focuses on the special features of the marine environment that affect the  
65 significance of the predicted environmental impacts and suggests actions that will advance understanding  
66 of these impacts.

67 *Key Words:* serious harm; deep-sea mining; seafloor massive sulphides; manganese nodules; cobalt-rich  
68 crusts; phosphorites

69

70 1. Introduction

71 Interest is accelerating in exploring the deep ocean with the intent to exploit seabed minerals [1].  
72 Minerals of interest include manganese nodules found on the abyssal plains [2], seafloor massive  
73 sulphides (SMS) found at active and inactive hydrothermal vents [3], cobalt-rich crusts on seamounts [4],  
74 and phosphorites found along continental margins [5]. Although commercial mining for deep-seabed  
75 minerals has yet to take place, all of these resources are under exploration in areas within and beyond  
76 national jurisdiction. Licenses to mine have been awarded for SMS exploitation by the government of  
77 Papua New Guinea, for metal-rich sediments in the Red Sea jointly by the Governments of Saudi Arabia  
78 and Sudan [6], and for offshore phosphorites (with environmental clearance pending) by the government  
79 of Namibia. Other companies and government agencies are submitting permit applications for  
80 exploitation within some national jurisdictions [7,8]. As the new industry of deep-seabed mining  
81 commences, regulatory bodies are faced with difficult permitting decisions, requiring balancing potential  
82 economic gains with impacts on other ocean users, local community and civil society concerns and  
83 international and national legal obligations to ensure effective protection of the marine environment  
84 from harmful effects that may arise from seabed mining activities [9].

85 Environmental protection regulations to be enacted and implemented by the ISA in the future, including  
86 functional distinctions and definitions of “harmful effects” and “serious harm,” will have far-reaching  
87 consequences both beyond and within national jurisdictions. Under UNCLOS, where mining activities may  
88 cause serious harm, the ISA has the power to: (i) set-aside areas where mining will not be permitted, (ii)  
89 deny a new application for a contract to conduct seabed mineral activities; (iii) suspend, alter or even  
90 terminate operations, and iv) hold the contractor and its sponsoring state liable for any environmental  
91 harm if it ensues (UNCLOS Art. 162((2) (w) and (x) and 165 (2)(k) and (l) and Annex III Article 18). Such  
92 standards will also inform national laws and regulations for mining activities within national jurisdiction,  
93 for such rules are to be “no less effective than” international rules, standards, recommended practices  
94 and procedures (UNCLOS Art. 208).

95 Of particular importance when designing a system to evaluate the significance of harm in the deep sea,  
96 where “serious harm” is used as the key trigger for preventive and precautionary action, are answers to  
97 the following questions:

- 98 1. How is “serious harm” defined in the context of deep-seabed mining?
- 99 2. What are the key factors or parameters to measure to inform the decision about whether an impact  
100 constitutes serious harm or not?
- 101 3. What are the special features of the deep-sea habitats targeted by mining companies that affect the  
102 significance of impacts?

103 In this paper, which is based on a workshop held in 2014, these questions are addressed by first  
104 examining current definitions that may inform our understanding of “serious harm” and the legal  
105 requirements to avoid such harm. Ecological and ecosystem parameters are considered that may be  
106 measured and there is a discussion of environmental thresholds and triggers for action when serious harm  
107 is predicted or is otherwise likely to occur. The mineral resources are then introduced, including  
108 distributions and ecological setting, the mining approach as understood at present, potential  
109 environmental impacts of mining, the distinctive environmental and ecological features of the associated  
110 ecosystem that inform the significance of impacts, and recommended actions to advance understanding  
111 of impacts. The need to consider cumulative impacts is presented, before finally concluding with the

112 overall implications of the issues that surround assessing the significance of harm for decision-makers and  
113 regulators with respect to deep-seabed mining activities. A key challenge will be to formulate regulations  
114 that prevent “serious harm” as well as ensure overall effective protection of the marine environment.  
115 Table 1 consolidates much of this information.

## 116 2. The international legal context for defining “Serious Harm”

117 The United Nations Convention on the Law of the Sea (UNCLOS) is the legal framework guiding  
118 international management of deep-seabed mining in the Area beyond national boundaries (the Area). In  
119 the interests of ensuring equitable, rational and sustainable development of seabed mineral resources,  
120 UNCLOS designates the seabed Area and its resources as the “common heritage of mankind” (UNCLOS,  
121 Part XI, Art. 136). (The term “resources” is defined in the context of UNCLOS Part XI on the Area as “all  
122 solid, liquid or gaseous mineral resources *in situ* in the Area at or beneath the seabed, including  
123 polymetallic nodules”). All rights to the resources of the Area are vested in mankind as a whole, on whose  
124 behalf the International Seabed Authority (ISA) is to act ((UNCLOS Art. 137.2). Activities in the Area are to  
125 be carried out for the benefit of mankind as a whole, taking into particular consideration the interests and  
126 needs of developing States (UNCLOS Art. 140). Developing States are to benefit through not only a share  
127 in the financial and other economic benefits derived from mining activities in the Area, but also through  
128 provisions designed to promote capacity building, technology transfer, and access to and participation in  
129 marine scientific research and mining-related activities in the Area (UNCLOS Art. 140, 143, 144, 148)  
130 including training programs conducted by the contractors (UNCLOS Annex III Article 15).[10]

131 An equally important objective as well as legal obligation under UNCLOS for both States and the ISA is to  
132 ensure “effective protection” of the marine environment from “harmful effects” which may arise from  
133 seabed mining activities (Article 145). For this purpose the ISA is required to adopt “appropriate rules,  
134 regulations and procedures for *inter alia*, (a) the prevention, reduction and control of pollution and other  
135 hazards to the marine environment, including the coastline, and of *interference with the ecological*  
136 *balance of the marine environment ...* and (b) the protection and conservation of the natural resources of  
137 the Area and the prevention of damage to the flora and fauna of the marine environment” (UNCLOS Art.  
138 145 (a) and (b))[11]. This is in addition to other obligations in UNCLOS that call for, *inter alia*, the  
139 protection and preservation of the marine environment,” and the taking of measures “necessary to  
140 protect and preserve rare or fragile ecosystems as well as the habitats of depleted, threatened or  
141 endangered species and other forms of marine life (UNCLOS Art. 192, 194(5)).

142 Existing ISA regulations for seabed mineral exploration of manganese nodules, SMS and cobalt-rich crusts  
143 provide only a definition for “serious harm”. Under these regulations, “serious harm to the marine  
144 environment” is defined to mean “any effect from activities in the Area on the marine environment which  
145 represents a **significant adverse change** in the marine environment determined according to the rules,  
146 regulations and procedures adopted by the Authority on the basis of internationally recognized standards  
147 and practices” (ISA Regulations (nodules), [12]; ISA Regulations (sulphides); [13], ISA Regulations (crusts),  
148 [14]). Such standards, as spelled out in the regulations and an Advisory Opinion by the International  
149 Tribunal for the Law of the Sea, are to ensure the application of “best environmental practices and the  
150 precautionary approach” ([13] Regulations 31(2); [15]).

151 The potential for serious harm entails serious consequences. As required by UNCLOS Art. 165(2)(I), the  
152 Legal and Technical Commission (LTC), the ISA’s advisory body, is to, among other tasks, develop  
153 recommendations to the Council, the ISA’s executive body, to disapprove mining in areas where

154 “substantial evidence indicates the risk of serious harm to the marine environment”. The LTC is also  
155 empowered to develop recommendations for emergency orders during mining operations to “prevent  
156 serious harm to the marine environment” (UNCLOS Art. 165 (k)). In turn, the ISA Council is required to  
157 issue emergency orders, which may include orders for the suspension or adjustment of operations, to  
158 prevent serious harm to the marine environment arising out of activities in the Area (UNCLOS, Art.  
159 162(2)(w)).

160 Unless mining proponents and permitting decision-makers have clear and comprehensive parameters for  
161 what constitutes both “effective protection” as well as “serious harm” and associated significant adverse  
162 change to the marine environment, there will be a risk that seabed mining could cause unacceptable  
163 impacts.

164 Some helpful guidance for defining serious harm may be drawn from the definition of “significant adverse  
165 impact” in the International Guidelines adopted in the context of deep-sea bottom fishing on the high  
166 seas by the FAO Food and Agriculture Organization in 2009. These guidelines were developed to help  
167 states and regional fisheries management organizations (RFMOs) implement a United Nations General  
168 Assembly Resolution of 2006 which called upon them to, among other things, “assess, on the basis of the  
169 best available scientific information, whether individual bottom fishing activities would have significant  
170 adverse impacts on vulnerable marine ecosystems and to ensure that, if it is assessed that these activities  
171 would have significant adverse impacts, they are managed to prevent such impacts, or not authorized to  
172 proceed” (UNGA Resolution 61/105 para 83(a)).

173 This FAO definition is particularly relevant as the ISA Mining Code contains a similar formulation with  
174 respect to exploration impacts, which provide that:

175 *“The Commission shall develop and implement procedures for determining, on the basis of the best*  
176 *available scientific and technical information....,whether proposed exploration activities in the Area*  
177 *would have serious harmful effects on vulnerable marine ecosystems [, including seamounts and*  
178 *hydrothermal vents], and ensure that, if it is determined that certain proposed exploration activities*  
179 *would have serious harmful effects on vulnerable marine ecosystems, those activities are managed*  
180 *to prevent such effects or not authorized to proceed.”* (ISA Regulations (nodules 31.4; sulphides  
181 33.4; crusts 33.4)

182 The FAO Guidelines provide that significant adverse impacts are “those that compromise ecosystem  
183 integrity” (FAO, 2009, para 17). It lists six factors to consider: (1) intensity and severity of the impact; (2)  
184 spatial extent of the impact relative to habitat availability; (3) sensitivity and vulnerability of the  
185 ecosystem to the impact; (4) ability for the ecosystem to recover; (5) the extent of ecosystem alteration;  
186 and (6) the timing and duration of the impact relative to species and habitat needs ([16] para 18). It  
187 further considers duration and frequency of impacts as metrics for determining significance ([16], para 19-  
188 20). In addition the authors recommend including the concepts of: (7) probability of impacts occurring;  
189 (8) cumulative effects of impacts, and (9) scientific uncertainty related to impacts, when determining  
190 what deep-seabed mining impacts should be considered “significant” The FAO Guidelines also provide  
191 criteria for identifying “vulnerable marine ecosystems” in the context of deep-seabed bottom fishing, but  
192 their applicability to seabed mining is beyond the scope of this paper.

193 In reality, assessing any changes to deep-sea ecosystems induced by mining activities is challenging at  
194 best. The remoteness and expense of studying these ecosystems has resulted in major knowledge gaps  
195 concerning habitat distribution (regionally and globally), ecosystem structure and function. These gaps

196 include species identities (most deep-sea species are undescribed), biodiversity, distribution patterns and  
197 biogeography, community distributions, dynamics, trophic relationships, population connectivity,  
198 physiological tolerances, ecosystem tolerances, and resilience.. Without this baseline information, it is  
199 difficult to assess the impacts of a human activity in space and time, to determine whether these impacts  
200 are enduring or transitory. The use of a systematic approach based on a robust ecological assessment of  
201 the key physical, biogeographic, ecological, and biodiversity features of the deep seafloor will be useful  
202 when dealing with the challenges of managing a large underexplored area [17]. Cumulative impacts of  
203 multiple mining actions (in space and time) and additive perturbations from direct human activities (e.g.,  
204 fishing activities, contaminants and spills), and climate-change related stressors (e.g., warming, ocean  
205 acidification and deoxygenation) must also be considered when evaluating the significance of [changes to](#)  
206 [and/or impacts on](#) deep-sea ecosystems [18,19].

### 207 3. Significant adverse change: Thresholds and triggers

208 An ecological threshold is a point at which changes in an important ecosystem property or phenomenon  
209 have exceeded normal ranges of variability [20]. Such thresholds may, but will not necessarily be,  
210 “tipping points” at which a small further change will abruptly produce a large ecosystem response [20]  
211 resulting in a regime shift (change in state). In the context of deep-seabed mining, ecological thresholds  
212 should help to inform the determination of when an adverse change and/or impact may be considered a  
213 significant one, i.e. ‘serious harm’. The identification of ecological thresholds requires, at the very least,  
214 knowledge of long-term (years to decades) average baseline conditions and natural ecological variability.  
215 Although natural variability is often determined from time series investigations of 3-25 years, the  
216 appropriate time period for assessment will be system-dependent [21]. With an understanding of  
217 ecological thresholds, decision-makers can determine: (1) what impacts are expected to exceed ecological  
218 thresholds and therefore should not be permitted; and (2) what impacts could exceed ecological  
219 thresholds and therefore require management, monitoring and then cessation of operations if the  
220 threshold is neared.

221 However, one of the greatest challenges for environmental management of the deep sea is the  
222 substantial lack of data, making the use of ecological thresholds for decision-making in deep-seabed  
223 mining a difficult one at best. The mandate to apply a precautionary approach and a lack of baseline data  
224 necessary to define ecological thresholds should lead to heightened restrictions, including at least slow  
225 ramping up of activities until thresholds are better characterized [1]. Key metrics that may serve as  
226 threshold indicators are measures of biodiversity, abundance, habitat quality, population connectivity,  
227 heterogeneity levels, and community productivity.

228 If information is not available to set particular ecological thresholds, a suite of other indicators can be  
229 used to determine the likelihood of significant adverse change and impacts, including those that address  
230 species-, community- or ecosystem-level impacts. Here all three ecological levels are considered.  
231 Significant species-level changes or impacts include: (i) extinction; (ii) significant decline in abundance; (iii)  
232 decline in foundation species; (iv) reduction below critical reproductive density; (v) loss of source  
233 populations; and/or (vi) loss of critical stepping-stone populations. Community-level impacts include (i)  
234 alteration of key trophic linkages among species in a community; (ii) reduction in species diversity beyond  
235 natural levels of variability; and/or (iii) regional declines in habitat heterogeneity, such as loss of entire  
236 habitats or community types. At the ecosystem-level, impairment of important ecosystem functions such  
237 as biomass production, nutrient recycling or carbon burial can lead to loss of major ecosystem services  
238 upon which society depends. They may include loss of carbon sequestration capacity, genetic resources,

239 or fisheries production. These impacts can be evaluated in local, regional or global contexts. While the  
240 concept of ecosystem services underlies many of the above indicators and metrics, threshold levels of  
241 decline in services have yet to be identified. These services are likely to vary by habitat, and the spatial  
242 and temporal scale at which changes are significant to the ecosystem have not been defined here.  
243 Additional measures that reflect key services are needed and a quantifiable measure of lost services will  
244 need to be incorporated into significance assessment [22].

#### 245 4. Deep-seabed mining resources and potential impacts

##### 246 4.1 Deep-seabed mining activities to date

247 As of 2016 there are 25 exploration contracts approved by the ISA: four for cobalt-rich crusts, each 3,000  
248 km<sup>2</sup> in the South Atlantic and Mid-Pacific; 16 for manganese nodules, each up to 75,000 km<sup>2</sup> in the Clarion  
249 Clipperton Fracture Zone in the Pacific (15) and in the Indian Ocean (1); and five for SMS, each  
250 approximately 10,000 km<sup>2</sup> in the Indian Ocean and on the Mid-Atlantic Ridge (Figure 1). There also has  
251 been commercial interest in deep-seabed minerals within national jurisdictions, including in the Pacific  
252 Islands region, Mexico, Namibia, New Zealand, Saudi Arabia, South Africa, and Sudan. For example, many  
253 deep-seabed mineral licenses for *exploration*— evaluating resources prior to the production phase of  
254 mining – have been issued over recent years for SMS by Fiji<sup>2</sup>, Papua New Guinea<sup>3</sup>, Solomon Islands<sup>4</sup> and  
255 Tonga.<sup>5</sup> In August 2015, the Cook Islands invited bids for deep-seabed mineral exploration of manganese  
256 nodules within its waters but none have been received.<sup>6</sup> The world’s first deep-seabed mining project may  
257 commence as early as 2019 in Papua New Guinea’s waters, although financial difficulties are causing  
258 delays.<sup>7</sup> The mining company, Nautilus Minerals Inc., was granted a mining lease in 2011 by Papua New  
259 Guinea Government in the South Pacific Bismarck Sea containing SMS deposits.<sup>8</sup> In 2010, Saudi Arabia and  
260 Sudan granted a production license for a SMS project known as ‘Atlantis II’ in the Red Sea, managed by a  
261 Saudi Arabian/Canadian consortium. However, there have been no public indications of imminent  
262 intention to commence production in this site.<sup>9</sup> The Namibian government has granted two mining  
263 licenses (to Lev Leviev and to Namibian Marine Phosphate), but a moratorium was instituted while further  
264 environmental impact assessment was conducted. As of August 2016 the official decision regarding  
265 marine phosphate mining in Namibian waters had not been announced, with the matter of  
266 strategically assessing the cumulative environmental impacts under review by the Government of  
267 Namibia.<sup>10</sup> As of Sept. 2016, a decision on phosphate mining in Namibian waters is still pending.  
268 Exploration and mining claims for phosphorites have been made in South Africa, New Zealand  
269 and Mexico, although the environmental ministries of the latter two States have recently rejected  
270 these based on environmental concerns<sup>11,12</sup>. It is important to note that no exploitation of deep-seabed  
271 minerals has taken place yet. This situation presents an unusual opportunity for the architects of relevant  
272 legal regimes, and the permitting decision-makers, to make informed decisions at the outset.

<sup>2</sup>[http://www.lands.gov.fj/images/tenementmaps/june15/DeepSeaExpl\\_June2015.pdf](http://www.lands.gov.fj/images/tenementmaps/june15/DeepSeaExpl_June2015.pdf)

<sup>3</sup><http://portal.mra.gov.pg/Map/> (search on ‘Nautilus’ and ‘Bismarck’)

<sup>4</sup> Page 14, SPC-SOPAC Division Published Report 151 – Lily, Tawake, Ishmael, 2012:

[http://www.sopac.org/dsm/public/files/reports/country/PR151\\_Solomon%20national%20workshop\\_hl-at.pdf](http://www.sopac.org/dsm/public/files/reports/country/PR151_Solomon%20national%20workshop_hl-at.pdf)

<sup>5</sup>[http://gsd.spc.int/dsm/public/files/2014/may/03\\_LepaolaVaea.pdf](http://gsd.spc.int/dsm/public/files/2014/may/03_LepaolaVaea.pdf)

<sup>6</sup><http://www.seabedmineralsauthority.gov.ck/cook-islands-seabed-minerals-tender-2015>

<sup>7</sup> <http://subseaworldnews.com/2016/06/29/nautilus-looks-to-save-its-deep-sea-mining-project/>

<sup>8</sup> <http://www.nautilusminerals.com/s/Projects-Solwara.asp> and news release of 16 June 2015, available here:

<http://www.nautilusminerals.com/s/Media-NewsReleases.asp?ReportID=712314>

<sup>9</sup> <http://www.diamondfields.com/s/AtlantisII.asp>

<sup>10</sup> B. Currie, pers. communication

<sup>11</sup> DECISION ON MARINE CONSENT APPLICATION Chatham Rock Phosphate Limited To mine phosphorite nodules on the Chatham Rise [http://www.epa.govt.nz/eez/EEZ000006/EEZ000006\\_CRP%20Final%20Version%20of%20Decision.pdf](http://www.epa.govt.nz/eez/EEZ000006/EEZ000006_CRP%20Final%20Version%20of%20Decision.pdf)

<sup>12</sup> [https://www.earthworksaction.org/earthblog/detail/victory\\_mexico\\_seabed\\_mining\\_project\\_scrapped#.V93Si07OrDI](https://www.earthworksaction.org/earthblog/detail/victory_mexico_seabed_mining_project_scrapped#.V93Si07OrDI)



273 Below four of the resources currently being considered for deep-seabed mining in national or  
274 international waters are addressed. Hydrothermal, sulfidic sediments (under license in the Red Sea) were  
275 not discussed at our workshop and thus are not treated here. The text below distinguishes **effective**  
276 **protection of the marine environment from** harmful effects, which is the general overarching objective,  
277 from “serious harm” that causes significant adverse impacts, and under UNCLOS as well as current  
278 regulations, should prevent mining from occurring in a specific area or require suspension or adjustment  
279 of mining activities to prevent serious harm.

## 280 **4.2 Manganese nodules**

281

282 **4.2.1 Mineral resource:** Manganese nodules (also called polymetallic nodules) are mineral precipitates of  
283 manganese, iron oxides and other metals. Nodules range in size from millimeters up to a half meter or  
284 more. The nodules of greatest commercial interest contain relatively high levels of nickel and copper (e.g.,  
285 1.5% of the nodule weight), cobalt, zinc, and traces of other metals (e.g., molybdenum, lithium) important  
286 to high-tech industries [23]. The nodules form extremely slowly (with estimated growth rates of 2-15 mm  
287 per million years) and occur over extensive areas in the abyssal Pacific, Indian and Atlantic Oceans, where  
288 they provide hard-substrate habitat for a variety of fauna (e.g., sponges, foraminifera). The nodules  
289 currently of greatest commercial interest occur beneath the relatively low productivity environment at  
290 10-20 degrees north of the equator in the Pacific Ocean in the Clarion Clipperton Zone (CCZ), within an  
291 area of ~6 million km<sup>2</sup>, with nodule abundances ranging from <1 to >35 kg m<sup>-2</sup> [23].

292 **4.2.2 Mining overview:** The extraction of manganese nodules is envisioned to be carried out by a series of  
293 remotely operated, technologically advanced nodule harvesters that are likely to plough, scrape, and/or  
294 vacuum the seafloor over large areas (300-800 km<sup>2</sup> of seabed per mining operation per year, [24]. As  
295 envisioned by some companies (e.g., UK Seabed Minerals), crushed or whole nodules and entrained  
296 sediments will be pulled up a riser pipe to the surface, where nodules will be offloaded to a production  
297 support vessel for transport to land. Sediment-containing water will most likely be returned to the ocean  
298 at the site at an, as yet, undetermined depth.

299 **4.2.3 Impacts from mining on habitat/resources:** Manganese nodule-mining operations could have major  
300 impacts over large abyssal regions. Removal of nodules will remove specialized fauna (e.g., foraminifera  
301 and sponges) that live on the nodules [25,26,27], together with organisms that live in the soft sediment  
302 patches between and under the nodules [24,28]. Epifaunal densities increase with nodule density, and  
303 alcyonacean and antipatharian corals are present in the CCZ only where nodules occur [29]. In addition,  
304 there appears to be a high diversity of nodule-obligate megafauna in the CCZ [30]. Nodule removal,  
305 sediment disturbance and plume perturbations have the potential to reduce habitat complexity,  
306 biodiversity and ecosystem function over large spatial scales both at the seafloor and in the water  
307 column. Some effects will likely persist for millennia because the formation of new nodules, and the  
308 habitats and heterogeneity they provide, is estimated to take millions of years [29].

309

310 The nodule extracting equipment will remove and disturb the top 15-40 cm of sediment that provides  
311 food for a high diversity of surface deposit-feeding organisms. The extracting equipment will likely also  
312 compress seafloor sediments beyond normal conditions, adversely affecting biota living within the  
313 sediments, benthic colonization and other processes (e.g., seafloor biogeochemistry; [31]. Nodule  
314 collection will cause sediment plumes during discharges that may disperse at least 10s and possibly  
315 hundreds of kilometers [24,27]. Such plumes have the potential to bury or smother seafloor organisms  
316 and habitats, and prevent larval settlement and colonisation, because background rates of sediment  
317 deposition are extremely low (e.g., [32,33]. Studies have shown negative effects from 1 cm of sediment  
318 deposition in the CCZ [34], and the most food-poor area of the CCZ (NW regions in the CCZ) will most  
319 likely be more sensitive to sedimentation. Plumes dispersing through the water column may clog the  
320 filtering membranes of suspension-feeding fauna (both benthic and pelagic). Resuspended sediments may  
321 also release oxygen-depleted pore waters and chemicals (e.g., heavy metals from the sediment) with

322 potential biogeochemical or ecotoxicological effects [35], and could affect vision, feeding and  
323 communication processes (e.g., bioluminescence) in the pelagic environment. The chronic effects of  
324 recurrent sediment plumes within an abyssal mining region remain unstudied but could be deleterious in  
325 regions such as the abyssal CCZ where natural sediment resuspension has not been documented and  
326 appears very unlikely based on observations and measurements [36].  
327

328 Waste materials from initial at-sea separation are expected to be released through a discharge pipe into  
329 the middle of the water column or at the water/seafloor interface, but the resulting sediment plumes may  
330 also cause impacts on midwater and mesopelagic species in the water column. Suspended particle effects  
331 on feeding, prey avoidance, and other ecological processes are likely but unstudied. In addition, biota  
332 throughout the water column might be affected by sediment leakage from the system used to lift the  
333 nodules to the surface, and sediment runoff from the mining support vessel could have local impacts on  
334 photosynthetic productivity in surface waters.  
335

336 In addition to physical impacts caused by mining, sound from mining machines, pumps, platforms and  
337 vessels may occur at the sediment-water interface, mid-depth water column and surface water column.  
338 Anthropogenic sound is known to cause harmful effects to marine mammals [37], but impacts on lower  
339 trophic level organisms are poorly understood, even in shallow-water environments. In addition to  
340 sound, light from mining operations may blind, attract, or misdirect organisms, altering their visual  
341 capabilities, communications, mate finding or prey avoidance capacity [38].  
342

343 **4.2.4 Special features that affect significance of predicted impacts:** Several traits characterize abyssal  
344 habitats where nodules are found that are key to understanding the significance of predicted impacts and  
345 recovery potential of affected ecosystems, and thus the potential for mining to inflict ‘serious harm’.  
346

347 (1) Manganese nodules are found underlying mesotrophic/oligotrophic water masses and the fauna  
348 inhabiting the water column and seafloor in these regions generally exist under extremely stable  
349 physical conditions (including low sedimentation rates, low current velocities and few/no resuspension  
350 events over decade to century time scale). **Therefore, extensive resuspension and deposition of  
351 sediments over large spatial scales will cause a substantial change to the existing ecosystem, with  
352 harmful effects. Some of these changes may be sufficiently widespread to constitute serious harm to  
353 vulnerable marine ecosystems and/or to the wider marine environment.**  
354

355 (2) Abyssal ecosystems in the CCZ are very food poor, with biomass, community structure, production,  
356 growth rates, and recolonization rates all controlled by the very low flux of particulate organic material  
357 sinking from the distant euphotic zone [39]. **Therefore, ecosystem recovery rates from mining  
358 disturbance will be very slow, so impacts on the sediment-dwelling biota from single mining  
359 activities may persist for decades to centuries, causing harmful effects.**  
360

361 (3) Manganese nodules provide much of the available hard substrate in the deep waters of the CCZ and  
362 other abyssal plains regions, and nodules can take millions of years to form [40]. **Nodule removal will  
363 eliminate much of this habitat, as well as the specialized nodule fauna** [29]. Surveys conducted in the  
364 CCZ along tracks from trawling or experimental mining simulations up to 37 years old suggest that  
365 epifauna is almost completely absent and recovery of the ecosystem in this area is slow [29]. **Complete  
366 removal of nodules will cause significant adverse change for long periods of time (millennia),  
367 possibly over areas hundreds of km in extent.**  
368

369 (4) Many of the fauna inhabiting soft sediments found amongst the nodules inhabit the top 5-10 cm of  
370 sediment [28], and their recovery from sediment disturbance is likely to take a long time, i.e., decades  
371 to centuries (e.g.,[41]. **Removal of sediment during nodule extraction will damage or kill the soft-  
372 sediment fauna over large spatial scales, reducing biodiversity and causing harmful effects within a  
373 mining claim. Timescales for recovery are poorly constrained, but small-scale disturbance tests  
374 suggest that recovery may take decades to centuries.**

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- (5) Much of the fauna inhabiting the abyssal regions where nodules occur has not been described or sampled. In those areas that have been sampled, levels of biodiversity appear to be high. Species and dispersal ranges are also poorly known, and the relationships between community structure and ecosystem function are uncertain. ***It is possible that the mining of manganese nodules could cause serious harm through the extinction of hundreds or more of undescribed species, especially those with small biogeographic distributions, thereby altering evolutionary potential, biodiversity (of species and genes), and ecosystem processes in the abyss. Such changes may be sufficient to be considered serious harm.***

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#### **4.2.5 Actions toward understanding significant impacts of manganese nodule extraction.**

- More extensive studies of pelagic and benthic biodiversity, species distributions and dispersal ranges, and ecosystem resilience, functions and services are needed in areas targeted for manganese nodule extraction to better characterize baseline conditions and assess potential significant mining impacts. Communication among scientific researchers working in the same region (e.g., on nearby claims) will be important.
- Realistic, large-scale mining disturbance studies may be needed to assess the spatial scales and intensities of disturbance resulting from mining (including cumulative impacts).
- Determining the time scales of recovery of affected soft-sediment and nodule communities and pelagic ecosystems is also required, but the very long recovery times expected (up to millennia) will make estimation of recovery times challenging.

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### **4.3 Seafloor massive sulphides**

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**4.3.1 Mineral resource:** Seafloor massive sulphides (SMS; also called polymetallic sulphides) are deposits found associated with active hydrothermal vents and inactive sulphide deposits in a variety of geological settings, including mid-ocean ridges, back-arc basins, volcanic arcs and intraplate volcanoes, at hundreds of highly localized vent sites, often at depths of 1200-3000 m [40,42,43]. Active vents are defined here as systems with evident hydrothermal fluid flux (i.e., temperature and/or chemical anomalies) supporting chemoautotrophic ecosystems typically dominated by invertebrate-microbial symbioses. Inactive sulphide deposits lack surficial evidence of hydrothermal fluid flux and lack dense populations of symbiont-hosting invertebrate taxa. Instead, they are often visually depleted in biota, with only an occasional megafaunal invertebrate observed. Because of this general perception of inactive sulphide deposits as biologically depauperate, they have not attracted much exploration or biological characterization. However, the Gorda Ridge (Escanaba Trough) sulphide mounds are one example of inactive deposits that host a large megafaunal population of brachiopods [44], sponges, corals and barnacles are reported in Manus Basin [45], and corals and echinoids are found in high density at inactive sites on the Kermadec volcanic arc [46].

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Sulphide deposits generated by seabed hydrothermal systems contain high-grade ore that includes minerals such as copper sulphide (chalcopyrite) and zinc sulphide (sphalerite), gold and silver [7]. The geochemistry of the host rock and fluid compositions in the different geological settings shape both the ore concentrations in the deposits and the composition and functioning of the ecosystem. At present, there are approximately 400 known active vent fields, with estimates of 1300 total on mid-ocean ridges and back-arc spreading centers [42,47,48]. Where these vent systems occur on intermediate- or slow-/ultra-slow spreading centers, they are expected to have accumulated fossil massive sulphide deposits that may have commercial potential.

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The distinction between active hydrothermal vents and inactive sulphide deposits is important in considering environmental management needs; the two types of ecosystems associated with these deposits support different communities and have different vulnerabilities to consider [49,50]. However,

423 prospective mine sites often contain both active and inactive deposits, and so the vulnerability of  
424 communities at both will need to be assessed.

425 **4.3.2 Mining overview:** Existing concepts for SMS mining envision the use of multiple remotely operated,  
426 large, technologically advanced machines to undertake open-pit mining once the sediment overburden is  
427 removed [51]. The machines will cut, crush, and gather the ore and then send it as slurry to the  
428 production support vessel via an enclosed riser and lifting system. Once on board the mining support  
429 vessel, the slurry will be dewatered to collect all but the finest particles (to ~10 µm), then seawater and  
430 fine particles will be discharged back into the sea at a depth close to the seabed. The ore will be  
431 transported to land for further processing. While single mining sites may cover local areas as small as 100  
432 m<sup>2</sup>; multiple sites in close proximity are likely to be mined, thus introducing cumulative harm within the  
433 exploited region that should be considered in assessing significant adverse impact.

#### 434 **4.3.3 Impacts from mining on habitat/resources**

435 **4.3.3.1 Active vents.** SMS mining operations will have a direct impact on the mining site, removing the  
436 substratum and its associated fauna and thus reducing diversity at all levels: genetic, species, functional,  
437 and habitat [38,43,52] and causing serious harm. The 3-dimensional structure of the habitat will be  
438 flattened, reducing habitat heterogeneity to a minimum and changing the substratum characteristics  
439 (e.g., porosity, particle size distribution, mineralogy), as well as the geochemical and hydrodynamic  
440 regimes [38,50]. High turbidity plumes with elevated metal concentrations generated by the mining  
441 activity and from the return into the sea of the water that has been separated from the ore material on-  
442 vessel [53] could affect pelagic and benthic populations downstream, potentially impeding vision,  
443 reducing bioluminescence, clogging organs of filter- and particle-feeders, disrupting larval development  
444 and settlement and potentially resulting in toxic effects from bioaccumulation of metals [27,43]. As with  
445 manganese nodule mining, light and noise from seabed activities may cause additional impacts. At active  
446 sites, these physico-chemical impacts will likely be temporary, since venting is expected to persist and  
447 new chimneys will precipitate and coalesce [54]. Local fauna will be crushed by the mining operations  
448 (sessile and slow moving fauna) or dispersed to other areas (mobile fauna such as fish and some large  
449 crustaceans), thus reducing biodiversity and abundance totally or to very low levels, in the absence of  
450 mitigation. This direct impact is particularly of concern for vent-endemic and rare species. At fast- and  
451 intermediate-spreading ridges, active vent communities are known to recover from catastrophic impacts  
452 (volcanic eruptions) on a decadal time scale [38]. The rate of recovery following a catastrophic  
453 disturbance at active vents on slow- or ultra-slow spreading ridges or in back-arc basins is unknown.

454  
455 **4.3.3.2 Inactive vents.** A single mining event at an inactive vent will eliminate local fauna and cause harm,  
456 but the extent to which this fauna is endemic is unknown. If the fauna of inactive vents is also found on  
457 hard substrata throughout a ridge system, the severity of the impact is considerably lessened in terms of  
458 lost biodiversity. However, where populations of particular taxa are relatively large at inactive vent sites,  
459 these populations could be disproportionately important for the maintenance of populations elsewhere  
460 through the supply of larval recruits. Recovery of inactive vents from mining activities is not known –in  
461 fact the fauna of inactive vents is hardly known [45, 49]– but is presumed to be slow, on the order of  
462 decades to centuries [38, 46]. To assess the potential for such recovery of populations and communities, a  
463 sound understanding of regional species distributions, their genetic diversity, reproductive ecology, and  
464 dispersal potential, as well as regional hydrodynamic regime that will drive dispersal of larvae, resulting  
465 gene flow and colonisation, is necessary. This must be placed within the context of temporal dynamics,  
466 natural variation, and succession of indicator species [52].

467  
468 At inactive sites, mining will remove the vertical topography, and modify seafloor texture and habitat  
469 heterogeneity; this will likely be a permanent effect at inactive sites, barring reactivation of fluid flow. If  
470 an inactive vent becomes active [55], once the ore is removed, the local physiography, biodiversity, and  
471 connectivity would be changed and could be enhanced.

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474 **4.3.4 Special features that affect the significance of predicted impacts at vents:**

475 **4.3.4.1 Active vents.** Because active vents are naturally dynamic (including becoming inactive and subject  
476 to destructive volcanic eruptions), the fauna exhibit fast growth rates and high reproductive output, that  
477 may make them resilient to major disturbances. However, other attributes may make them particularly  
478 susceptible to cumulative impacts of multiple mining events in a region and have led to their listing as  
479 indicators of Vulnerable Marine Ecosystems [16]. Several characteristics distinguish active hydrothermal  
480 vents from most other habitats in the marine biome [56, 57, 58] and are key to understanding the  
481 significance of predicted impacts and recovery potential of affected ecosystems, and thus the potential  
482 for mining to inflict a significant adverse impact and ‘serious harm’:

- 483 (1) Active hydrothermal vents are spatially very limited (on the order of 100 to 500 m maximum  
484 dimension) and distributed linearly along mid-ocean ridges or patchily in seamount provinces. The  
485 population of a species at a given vent is part of a larger metapopulation. **The island-like distribution  
486 of vent habitats means that a given metapopulation may be susceptible to cumulative mining  
487 impacts (e.g., loss of reproductive populations) that interfere with connectivity among populations.  
488 A break in connectivity (e.g., loss of populations at multiple and/or critical habitats) would result in  
489 isolation, loss of biodiversity, and other ecological consequences, sufficient to cause significant  
490 adverse impacts and serious harm to the marine environment.**
- 491 (2) Active hydrothermal vents host taxa that are deemed likely to be adapted for and restricted to the  
492 vent environment, that is, the taxa may be endemic to vents. **Thus vent-endemic taxa have limited  
493 habitat available to them and could be at risk of a significant adverse impact and serious harm and  
494 (including extinction) if multiple and/or critical habitats are lost** beyond natural levels of habitat loss (  
495 i.e., single sites becoming inactive)..
- 496 (3) Geothermal activity of vents serves as a source of reduced compounds that fuel microbial  
497 chemoautotrophic primary production, which in turn supports higher trophic levels, and has an  
498 ecological influence beyond the boundaries of the active vent [59]. **Disruption of primary productivity  
499 at active vents may have a harmful effect on the secondary productivity of the surrounding seabed  
500 environment. There is no evidence to date to suggest that this would be a significant adverse  
501 impact.**
- 502 (4) Dominant host-invertebrate populations at active hydrothermal vents form biogenic habitats that  
503 support abundant smaller taxa. **Loss of biogenic habitats at an active vent may place other taxa at  
504 risk, particularly if during the recovery phase, community structure shifts to an altered state that  
505 does not include the lost biogenic habitat. This habitat loss could result in a significant change  
506 although whether the impact is adverse or positive is difficult to predict without further knowledge.**
- 507 (5) Plumes generated by mining activities will be enriched in toxic metals. **If these toxic metals become  
508 bioavailable, they could have a harmful effect on the biota (pelagic and benthic) in the plume  
509 shadow, and through bioaccumulation. The extent to which this bioavailability of toxic metals could  
510 become a significant adverse impact is not known.**

511 **4.3.4.2 Inactive vents:**

- 512 1) Inactive sulphide deposits are very poorly characterized. **Particularly if there should prove to be a  
513 rare fauna endemic to this environment, mining activities have the potential to cause a significant  
514 adverse impact, including extinction of taxa.**
- 515 2) Inactive sulphide deposits may be food poor, with biomass, community structure, production, growth  
516 rates, and recolonization rates all controlled by the very low flux of particulate organic material  
517 sinking from the distant euphotic zone [39]. **Therefore, ecosystem recovery rates from mining  
518 disturbance will be very slow, so impacts to the sediment-dwelling biota from single mining  
519 activities may persist for decades to centuries, causing harmful effects and a significant adverse  
520 impact.**

- 521 3) Inactive vent sites of interest for mining may exist in close proximity to (or be interspersed with),  
522 active vent sites, and fauna may benefit from the higher productivity associated with the active  
523 venting. This increased food availability may support relatively large populations of filter/suspension-  
524 feeding fauna such as corals, which are often long-lived. **Therefore, community recovery rates at**  
525 **inactive sites may be long (e.g., hundreds of years), and mining disturbance could also cause**  
526 **significant adverse change to the connectivity of those taxa that rely disproportionately on larval**  
527 **supply from large populations at inactive vent sites.**
- 528 4) Plumes generated by mining activities will be enriched in toxic metals. **If these toxic metals become**  
529 **bioavailable, could have a harmful effect on the biota (pelagic and benthic) in the plume shadow**  
530 **and through bioaccumulation. The extent to which bioavailability of toxic metals could become a**  
531 **significant adverse impact is not known.**

532

533 **4.3.5 Actions toward understanding whether there will be significant adverse impacts of SMS extraction**  
534 **at active hydrothermal vents and inactive sulphide deposits.**

- 535 • A thorough knowledge of the regional distribution of active and inactive sites and their  
536 associated fauna is necessary to understand regional variability in community composition and  
537 identify potential source and sink populations.
- 538 • Understanding connectivity of populations at active and inactive sites is critical to understanding  
539 whether mining and cumulative mining impacts in a region will result in a significant adverse  
540 effect (e.g., loss of biodiversity), in addition to serious harm. A related knowledge gap to fill is  
541 the extent to which local populations are maintained through local or long-distance recruitment  
542 events.
- 543 • Better understanding of natural community variability, succession patterns, potential alternative  
544 states for active hydrothermal vents are key ecological dynamics that would help in assessing  
545 likelihood of significant adverse effects.
- 546 • For inactive sulphide deposits, a thorough characterization of the colonizing fauna and their  
547 ecological attributes, including degree of endemism, growth rates, fecundity, and recruitment, is  
548 critical to understand the potential for causing a significant adverse impact through mining  
549 events.
- 550 • Understanding the ecotoxicology of plumes generated by seabed mining activities on pelagic and  
551 benthic biota in the plume shadow.

552 **4.4 Cobalt-rich crusts**

553 **4.4.1 Mineral resource:** Cobalt-rich crusts (also called polymetallic crusts) form a thin mineral layer  
554 through precipitation over millions of years from the surrounding seawater, and accumulate on  
555 hardground areas of seafloor swept clear of sediments by current flow [40,60]. The thickest deposits (up  
556 to 25 cm) occur on the summits and flanks of seamounts (especially large flat-topped guyots), ridges and  
557 plateaus, fused to the basal rock. These features occur in all oceans of the world, but are most common in  
558 the Pacific Ocean, where there are estimated to be at least 50,000 seamounts and knolls [61]. The main  
559 constituents of the crusts are iron and manganese, although many other minerals occur in smaller  
560 amounts. The metals of commercial interest in cobalt-rich crusts are cobalt, nickel and manganese, and  
561 are most enriched at water depths of 800 to 2500 m [62]. The crusts also contain rare-earth elements [40,  
562 63].

563 **4.4.2 Mining overview:** The mining of cobalt-rich crusts is more technologically complex than for  
564 manganese nodules (which occur on/in soft sediments) or SMS (which protrude from the seafloor in  
565 brittle structures) [51]. Although it will similarly require the use of large, remotely operated,  
566 technologically advanced machinery to dig into and cut, crush and gather the ore, and send it as slurry to  
567 the production support vessel through a riser and lifting system, or as whole rock material in a chain of  
568 bucket-type containers. The variable thickness of the crusts combined with the steep and rugged  
569 seamount terrain makes design and operation of the collection tools difficult. However, once on board  
570 the vessel, the processes are similar to the other mineral types. Slurry will be dewatered, the seawater

571 and discard products will be discharged back into the sea and the ore transported to land for processing  
572 [51]. A mining-site model developed by Hein et al. [64], and endorsed by the ISA in its cobalt-rich crust  
573 regulations, involves a combination of 20 km<sup>2</sup> sites and a total area of about 260 km<sup>2</sup> over a 20-year  
574 mining life span. Subsequently, He et al. [65] suggested that to be profitable a mine site would need to  
575 cover a total area of 1,214 km<sup>2</sup>, and be mined for 20 years to produce 1 million wet tons. This form of  
576 mining would require multiple mining sites on multiple seamounts in close proximity, and would cause  
577 cumulative impacts within the exploited region.

578 **4.4.3 Impacts from mining on habitat and resources:** Extraction of cobalt-rich crusts will cause impacts at  
579 the seafloor, through the water column, and potentially at the ocean surface [66]. The most direct and  
580 substantial effects will be on the habitats and benthic fauna at the seafloor. Here there will be substantial  
581 physical alteration of the seafloor as the crust is removed, the overall relief of the surface of the  
582 seamount will be flattened to an extent, and the amount of soft sediment will increase. Hence there is  
583 expected to be reduction in habitat heterogeneity, and changes in the geochemical characteristics of  
584 seafloor sediments. Seamounts often have high diversity and density of large, sessile animals, such as  
585 sponges and corals, as well as giant protozoans called xenophyophores, which form a biogenic habitat for  
586 other communities [67,68,69]. These organisms will be affected directly by the mining operations, and will  
587 not survive. Mobile animals, such as fish and crustaceans, may be able to disperse to other areas, but the  
588 overall biodiversity at the mining site will be reduced to very low levels. These impacts are particularly  
589 important for endemic and rare species. The capacity of affected communities to recover may be low.  
590 Typically, deep-sea invertebrate and fish species are long-lived and have slow growth rates, and their  
591 ability to recover from human disturbance (such as fishing or mining) is very low [70,71]. The potential for  
592 recovery will be affected by substrate changes, with soft sediment from mining plumes being unsuitable  
593 for many sessile species, and potentially smothering small animals and clogging the feeding structures of  
594 suspension feeders. The availability of source populations on nearby seamounts that will provide the  
595 necessary propagules (larvae, juveniles and dispersing adults) is poorly understood. Hence, both physical  
596 and biogenic habitat will be severely impacted over a long period of time (likely many decades to  
597 centuries for the fauna).

598  
599 Mining for cobalt-rich crusts will generate plumes of sediment, both from the physical disturbance of the  
600 seafloor, and from any discharge of processing waste. This sediment will have direct impacts on benthic  
601 communities through smothering and burying of animals, clogging of feeding structures, preventing larval  
602 settlement and colonisation, and indirectly through metal release and accumulation through the food  
603 chain. The vigorous hydrodynamic regime on seamounts suggests that the “downstream” extent of  
604 sediment plume impacts could reach well beyond the direct site of mining, over 100s of meters [72].  
605 Depending on the discharge depth of waste water (including fines, and small rocks), there can be impacts  
606 on plankton and fish communities throughout the water column. Effects include: potential oxygen  
607 depletion; nutrient and trace metal enrichment; changes in water clarity affecting visual predators;  
608 behavioural changes of plankton, mesopelagic fish and marine mammals from the plume density,  
609 contaminant composition, and associated noise; bioaccumulation of toxic metals in higher predators;  
610 toxic effects for early life history stages (e.g., larvae), and direct mortality of small plankton and  
611 mesopelagic fish [73].

612  
613 **4.4.4 Special features that affect significance of predicted impacts:** Limited biological survey work has  
614 been directed specifically at cobalt-rich crust habitats, but many studies have occurred on seamounts that  
615 are known to have such crusts. From these a number of key biodiversity characteristics can be defined  
616 (after [66,74,75]) that are important for understanding the significance of impact and recovery potential of  
617 seamount ecosystems.

618  
619 (1) The dominant benthic fauna on seamounts are often sessile corals and sponges, which can form  
620 dense reef-like structures [76], and be much more abundant on seamounts than on continental  
621 slope habitats (e.g., [77]). The complexity and fragility of these taxa make them highly  
622 vulnerable to impact by mining operations. These corals and sponges provide important food,

623 habitat and refuge for a large number for associated invertebrates and fish species (e.g.,  
624 [78,79,80]. **Hence loss of the main components of many cobalt-rich crust benthic communities**  
625 **will have a follow-on effect to other associated species, causing serious harm through**  
626 **significant adverse change.**

627 (2) The key coral and sponge species are filter or suspension feeders, which rely upon ingesting  
628 small particulate matter from the passing seawater. High sediment/particle loading from a  
629 sediment plume can clog polyps and feeding pores, causing a reduction in respiration, or death  
630 of the animal. **Harmful effects from mining operations will extend over the dispersal distance**  
631 **of the plume, i.e., over a much larger area than the direct mining site. Whether this causes**  
632 **serious harm will vary depending on the depths and current flow characteristics of each**  
633 **seamount where mining is conducted.**

634 (3) Many deep seamount species have high longevity, and slow growth-rates [57,68,76]. Seamount  
635 fish, which comprise major fisheries, may be one hundred or more years old, while deep-sea  
636 corals can be several thousands of years old (see review in [71]). Studies at the bathyal depths  
637 of cobalt-rich crusts have examined changes following cessation of trawl fishing and indicate  
638 that after up to 10 years there was no evidence of stony coral regrowth [70]. **Recovery of**  
639 **community composition, biomass and biodiversity levels and functioning similar to the**  
640 **original state is likely to be very slow in part due to great longevity of organisms. Significant**  
641 **adverse change in long-lived seamount species will constitute serious harm, in part due to**  
642 **their biodiversity- and fisheries-support functions.**

643 (4) Recent work has demonstrated that seamounts with differing levels of cobalt-rich crusts exhibit  
644 high faunal variability [74, 75, 81], with variations in species composition and abundance  
645 between cobalt-rich crust seamounts [74]. **The differences in species composition and**  
646 **abundance between seamounts means the impacts of mining will vary with location, and**  
647 **cannot be assumed to be the same even for adjacent mining sites.**

648 (5) Faunal composition changes with depth, and a potential mining depth range of 800–2500 m  
649 would impact a wide range of communities with differing species composition and abundance  
650 on a single seamount [74]. The vulnerability of communities will also change. **The depth range**  
651 **covered by a mining project must be considered when evaluating whether significant adverse**  
652 **changes will occur.**

653 The food chain is supported by primary productivity based on energy from photosynthesis. The  
654 topography of a seamount can cause upwelling of nutrients and localised circulation that traps  
655 plankton [82, 83], but support of high productivity at seamounts is believed to be largely through  
656 pelagic food webs, via horizontal advection of plankton from surrounding areas (meaning pelagic  
657 energy sources are continuously renewed) and by the seamount trapping vertically migrating  
658 plankton [68,82,84]. Chemosynthetic sources of production are rare but may be associated with  
659 volcanic activity on some seamounts with cobalt-rich crusts. Hence, although direct physical  
660 impact on the seafloor will affect production of benthic invertebrates, the greater risk is any  
661 effect from near seafloor and midwater sediment plumes on the pelagic food source and  
662 production-especially the return processing waste plumes which might be higher in the water  
663 column. **Mining activities, if the return sediment plume effects are controlled, might not alter**  
664 **overall seamount productivity levels to a significant extent.**

665 (7) Many benthic species are reported to be endemic to a seamount or group of seamounts. Whilst  
666 this is partly an artifact of sampling [85]. Stocks and Hart [86] found, overall, about 20% of seamount  
667 species had a restricted distribution. **Mining impacts that result in the loss of certain endemic**  
668 **species (i.e., those with restricted geographic distributions) would represent a significant adverse**  
669 **change causing serious harm.**

670

#### 671 **4.4.5 Actions toward understanding significant impacts of cobalt-rich crust extraction.**



- 672 • Given the depth range of cobalt-rich crust habitat, interactions with, and cumulative impacts  
673 from, commercial deep-sea fisheries and long-term ocean acidification need to be considered.  
674 • Improved knowledge is needed of the composition, structure and function of settings where  
675 cobalt-rich crusts are common, sensitivity of benthic fauna to changes to the substrate texture  
676 (e.g., hard rock to soft sediment), and regional population connectivity (e.g., sources and sinks).  
677 • The effects of sediment plumes generated by crust mining are poorly understood, and there is a  
678 pressing need for a greater understanding of increased sediment loads, and any associated  
679 ecotoxicity, on benthic and benthic-pelagic fauna.

## 680 **4.6 Phosphorites**

681 **4.6.1 Mineral resources:** Phosphorites (also called phosphates) are widely distributed in the marine  
682 environment, mainly in outer shelf and upper slope unconsolidated sediments on continental margins,  
683 and particularly along boundary upwelling areas where their biogenic origin is linked to organic-rich  
684 surface sediments [86, 88]. Rich phosphorite deposits are found in the Benguela, Humboldt, California  
685 and Canary upwelling systems, off the west coast of India, and off the southern and western shelf of  
686 South Africa. Other phosphorite deposits not linked to present-day upwelling (but including regions of  
687 high productivity) are found off the east coast of North America, off Australia and New Zealand and at  
688 several island sites in the Pacific and Indian Oceans [5,85,89,90]. The character of the deposit varies with  
689 locality, from small granules <500 µm, to lamellae, crusts, concretions and nodules several centimeters in  
690 diameter. Deposits are concentrated at varying depths within surface sediments.

691 Phosphorites contain significant amounts of phosphate resulting from the concentration of the mineral  
692 apatite or carbonate fluorapatite. Bacteria, under low oxygen conditions, are thought to play a key role in  
693 the formation of modern and ancient phosphorite deposits [91, 92, 93, 94]. Changes in sea level and  
694 winnowing by currents have concentrated deposits into areas now being considered for mining. The  
695 unique combination of changing sea levels over geological time and the high surface productivity due to  
696 upwelling off the Namibian coast, for example, resulted in the formation of large areas of phosphorus-rich  
697 sediments [95, 96].

698 **4.6.2 Mining overview:** A perceived future scarcity of terrestrially mined “rock phosphate” to  
699 manufacture phosphate-based fertilizer led to the present interest to mine phosphate from the ocean.  
700 The targeted deposits occur on the continental margins (both within the EEZs and extended continental  
701 shelf) of the host coastal States, in water depths varying from 50 m to 900 m. Recovery of the  
702 phosphorite will require bulk removal of the sediments just beyond the depths of maximum  
703 concentration, which can vary by locality (e.g., off New Zealand mining has been proposed to a sediment  
704 depth of at least 50 cm, and proposed to 3 meters deep off Namibia). To date, wide-head suction  
705 dredging of the seabed is proposed to pump tracts of sediment in bulk into large dredge-hopper vessels.  
706 Sorting the granules/nodules from the bulk is proposed in various ways: either vessel-transport of the  
707 entire bulk to coastal, land-processing facilities, or if phosphorites are large enough (i.e., if majority are  
708 nodules), sieving and sorting will take place onboard to collect the phosphorites, with release of  
709 unwanted bulk sediment back into the sea either at the sea surface or more likely at depth via a sinker  
710 pipe and diffuser. Some proposed mining includes further processing to fertilizer at the coastal sites, with  
711 effluent disposed into the sea.

712 **4.6.3. Impacts from mining on habitat/resources:** There have been a limited number of studies of the  
713 impacts or potential impacts of phosphorite mining and processing on the marine environment published

714 in the scientific literature (e.g., [97, 98, 99]). However, recent interest in phosphorite deposits off Namibia,  
715 South Africa, New Zealand and Mexico has resulted in a large number of studies assessing the various  
716 potential impacts resulting from phosphorite mining, which are published as EIA reports (Namibia -  
717 <http://www.namphos.com/>, New Zealand - [http://www.epa.govt.nz/EEZ/chatham\\_rock\\_phosphate](http://www.epa.govt.nz/EEZ/chatham_rock_phosphate)). An  
718 overarching impact is the total removal of large areas of seabed for continuous mining effort over many  
719 years (e.g., 30 km<sup>2</sup> annually over at least 15 years has been proposed off New Zealand). This scale of  
720 mining is likely to have more important long-term ecosystem impacts than recognizable short-term point  
721 impacts. For phosphorites found in upwelling regions in organic-rich suboxic to anoxic sediments,  
722 excavation and exposure of such sediments pose a special set of risks to these ecosystems.

723 The most significant impacts from phosphate mining are likely those associated with: (i) large-scale  
724 excavation and removal of soft sediment and hard substrates (e.g. large nodules on sediment surface),  
725 affecting benthic community composition and abundance, as well as recovery rates of sediment  
726 stratification, sediment biogeochemistry and rates of faunal repopulation; (ii) long-term and continual  
727 sediment plumes, both from mining excavation and processing, that could adversely affect ecosystem  
728 functioning of both benthic and pelagic components; (iii) disruption of important fish areas: not only  
729 fishing grounds but breeding, spawning and nursery areas of both commercial and non-commercial  
730 species; (iv) contamination of coastal waters from both a) exposure of deep anoxic and metal-rich  
731 sediments to the overlying water column and b) processing effluent returned to the sea; and (v) the  
732 possibility of large-scale sediment removal and redistribution leading to tipping points regarding  
733 tolerances of both benthic and pelagic sea life to physical and chemical parameters such as dissolved  
734 oxygen, turbidity, and organic matter content of the sediment which could lead to ecosystem regime  
735 shifts. There is significant concern about displacement of species of commercial demersal fish leading to  
736 food-web disturbances [100] and combined impacts on food-safety levels for consumption and marketing  
737 resulting from release of contaminants (e.g., cadmium, uranium) from sediments and disposal of  
738 processing waste.

739 **4.6.4 Special features that affect significance of predicted impacts:** Management of phosphorite mining  
740 may require different considerations from the management of other seabed mining because most  
741 phosphorite deposits fall within the exclusive economic zones (EEZs) or extended continental shelves of  
742 States, and are therefore governed under national law. These deposits fall in areas that are close to shore  
743 and are subject to sometimes intense use by humans for fishing, oil and gas extraction, shipping and by  
744 endangered marine mammals, turtles and seabirds for migration, mating, feeding and reproduction. Thus  
745 these areas may contain more ecologically sensitive species and habitats and may be more susceptible to  
746 cumulative impacts than those discussed previously [101].

- 747 (1) Continual large-scale mining activities will be necessary to make mining operations economical.  
748 ***This scale of activity will disturb large areas of the seafloor and could generate a continual***  
749 ***sediment plume with multiple potentially harmful effects throughout the water column, which,***  
750 ***if severe, could lead to serious harm through wholesale, significant adverse ecosystem change.***
- 751 (2) Mining areas coincide with many highly productive areas (e.g., upwelling regions) that are of high  
752 conservation value for marine mammals, seabirds, and sensitive habitats such as deep-water  
753 coral and sponge beds that may be protected by law. ***Damage to organisms or habitats of high***  
754 ***conservation status and with valuable ecosystem services may represent significant adverse***  
755 ***change causing serious harm.***
- 756 (3) Highly productive areas where phosphorite deposits are found are often areas of high priority for  
757 other commercial activities such as fishing. ***Therefore, there is direct conflict with other***

- 758 *commercial activities that rely on healthy ecosystem functioning and the potential for*  
759 *cumulative effects that cause serious harm, resulting from at least two forms of disturbance.*
- 760 (4) Upwelling areas, and other areas of high productivity, are acknowledged as highly variable  
761 environments. **Natural high spatial and temporal variability make it difficult to determine**  
762 **significant adverse change without extensive baseline data and sophisticated monitoring in**  
763 **these areas, which are lacking currently.**
- 764 (5) Phosphorites can occur in upwelling areas of low oxygen (e.g., the margins of Namibia, South  
765 Africa, Peru or Mexico) where the ecosystems are already stressed. **Low oxygen levels slow**  
766 **ecosystem recovery following disturbance. Increased oxygen demand from mining activities**  
767 **could cross tipping points for biota in the area, enhancing the likelihood of serious harm**  
768 **through significant adverse impacts from mining activity.**
- 769 (6) The low-oxygen, high-sulphide conditions associated with some phosphorites also host an array  
770 of microbes with genetic potential to function on under extreme conditions. **Mining impacts may**  
771 **cause loss of genetic diversity with unknown biotechnology value.**

772

#### 773 **4.6.5 Actions toward understanding significant impacts of phosphorite extraction.**

- 774 • Additional information is needed on microbial and faunal composition, connectivity, temporal  
775 dynamics and recovery potential of the biota associated with phosphorites of different sizes and  
776 textures. Both soft-sediment and hard-substrate faunas, and their ecological functions and  
777 ecosystem services require characterization.
- 778 • Given the proximity to coastal activities, interactions with, and cumulative impacts from, other  
779 ecosystem services associated with commercial fisheries, genetic resource potential, endangered  
780 and threatened species migration and habitation need to be addressed as a priority, as these  
781 provide direct vital services to the States concerned.

#### 782 **4.7 Significant effects: Cumulative impacts.**

783 Mining site impacts may be multifaceted, resulting in cumulative impacts in a setting or region, where the  
784 effect of more than one stressor can result in a magnified impact than the stressors taken individually  
785 [18]. Knowledge of cumulative impacts will inform the determination of serious harm across multiple  
786 sectors, and is essential to development of strategic environmental assessments and management plans.  
787 Cumulative impacts can occur at many levels: (i) multiple mining operations (by one or more contractors)  
788 within a sector and region (such as in the CCZ); (ii) multiple impacts from different mineral resource  
789 sectors; and (iii) overlapping impacts from non-mining sectors that coincide with mining impacts (e.g.,  
790 fisheries), including stressors related to climate change and pollution. Evaluation of cumulative impacts  
791 should include past, present and reasonably foreseeable future impacts. Estimating the magnitude of the  
792 impact of cumulative mining activities within a region is difficult with current knowledge, in part because  
793 of limited knowledge of the scales over which ecosystem structure and function play out. Unknowns  
794 include for example, the size, composition and distribution of source populations that will provide the  
795 larvae, juveniles and motile adults and dispersal potential of the source organisms available for  
796 recolonization and population recovery [102].

797 An additional challenge is knowing the nature and extent of all relevant activities and sources of change  
798 other than the targeted mining activity. The impacts from non-mining sectors, including oil and gas  
799 extraction, tourism, fishing, shipping, submarine cables, waste disposal, marine litter, chemical pollution,  
800 natural products, or research may be deemed acceptable individually, but when taken together with deep  
801 seabed mining can create significant impacts. For example, phosphorite lease claims in the national

802 waters of Namibia and Mexico intersect key habitat or nursery grounds essential to local fishes, while off  
803 New Zealand the phosphorite claims overlapped a marine protected area, one of a number of closed  
804 areas instigated by the fishing industry [103]. Another example comes from Papua New Guinea where  
805 submarine tailings placement from terrestrial mining [104] may potentially interact with those of seabed  
806 mining for SMS. Although the locations and water depths of seabed mining and other activities may not  
807 coincide, the physical transport of plumes and ontogenetic migrations of larvae and juveniles might cause  
808 cumulative impacts on some deep-sea populations.

809 Finally, ocean acidification and climate change operate at both global and regional scales and can affect  
810 recovery and resilience of many of the ecosystems treated here [105, 106]. Stress from warming  
811 temperatures, ocean acidification, and ocean deoxygenation as well as altered particulate organic carbon  
812 flux or circulation patterns may differentially affect hydrothermal vent, seamount, abyssal and continental  
813 margin systems depending on their location and water depth [19, 39].

#### 814 4.8 Conclusions and implications.

815 The paper authors have sought to understand what deep-seabed mining impacts have the potential to  
816 cause serious harm to the marine environment, harm that may among other things, invoke decisions not  
817 to mine in a specific area, not to approve a particular application for a contract to mine, to suspend or  
818 stop mining, to require adjustment of operations to avoid serious harm, and/or to provide financial  
819 compensation if harm ensues. Such an understanding should also indicate the type of impacts that mining  
820 operations should be designed to avoid (together with other proactive regulations to ensure effective  
821 protection).

822 While there are differences in extraction technology and methods used between different deep-seabed  
823 mineral types and projects, seabed mining actions that may cause harmful effects or serious harm across  
824 all targeted resources include: direct removal and destruction of seafloor habitat and organisms;  
825 alteration of the substrate and its geochemistry; modification of sedimentation rates and food webs;  
826 changes in substrate availability, heterogeneity and flow regimes; suspended sediment plumes; released  
827 toxins; and contamination associated with noise, light or chemical leakage during the extraction and  
828 removal processes. These impacts are expected to occur in benthic communities across hard and soft  
829 substrates and in the pelagic realm. Table 1 summarizes how these impacts relate to the potential for  
830 significant adverse change causing serious harm at sites where mining may take place for each of the  
831 mineral resources considered here.

832 However, there are clearly major knowledge gaps and uncertainties and these impel invocation of the  
833 precautionary approach. The application of this approach could include a clear requirement that:  
834 “Activities in the Area shall only take place if they do not cause serious harm to the marine environment”,  
835 the standard envisaged by the drafters of the first set of mining regulations in 1990 ([107] article 2(2)).

836 To construct rules, regulations and procedures capable of ensuring effective protection of the marine  
837 environment from the harmful effects of mining activities and avoiding serious harm, a well-defined  
838 understanding of what may or may not constitute significant adverse change in deep-sea biodiversity as  
839 well as ecosystem structure and function will be needed. Such an understanding is required prior to the  
840 onset of commercial-scale mining operations to prevent long term, potentially irreversible harm. An  
841 ability to identify and quantify significant effects, and the valuation of the key ecosystem services with  
842 which they are associated, will be necessary to implement appropriate environmental impact

843 assessments, environmental management plans and other regulations and payment regimes (including  
844 environmental damage assessment) associated with deep-seabed mining [22].

845 Scientific understanding about the impacts of mining will need to improve, and there are a number of  
846 national and international efforts currently underway which aim to achieve this goal, including the EU-  
847 funded MIDAS project (Managing Impacts of Deep Sea Resource Exploitation), among others. Disturbance  
848 experiments conducted in nodule provinces, on seamounts and at hydrothermal vents have to date  
849 provided valuable but limited insight into colonization and recovery rates in these realms because these  
850 relatively small-scale, low-intensity studies cannot replicate or predict mining impacts associated with  
851 exploitation, which will occur at much larger spatial scales and intensities. Nonetheless it is important to  
852 conduct both *in situ* and laboratory experiments in order that our understanding of serious harm from  
853 significant adverse change induced by deep-seabed mining improves, and ecological thresholds can be  
854 identified for use by regulatory authorities. Some States that have mineral resources of interest to  
855 companies do not yet have legal regulatory frameworks or official competency to assess or regulate  
856 marine mining (for example for phosphorites or SMS) within their EEZs. Legislation and regulatory  
857 conditions need to be instituted to complement those for other established industrial activities such as  
858 fishing.

859 While this article focuses on deep-seabed mining for minerals, these guidelines for identifying harmful  
860 effects and serious harm from potential significant adverse change may also help with management of  
861 future energy or bioprospecting interests in deep-seabed resource extraction. For example, there is great  
862 interest in exploring and exploiting novel chemical compounds and structures from deep-sea organisms  
863 that may prove useful for pharmaceuticals, molecular probes, enzymes, cosmetics, nutritional  
864 supplements, agrichemicals [108], bioinspired materials [109] and even climate remediation [110]. In the  
865 future, marine mining of methane hydrates as a novel energy source may also be of commercial interest  
866 [111]. Each of these commercial developments could have some adverse impacts on deep-sea ecosystems  
867 in common with those identified above, and some unique to the resource and its setting, depending on  
868 scale and intensity.

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881 References

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- 883 1. Wedding, L.M., S.M. Reiter, C.R. Smith, K.M. Gjerde, J.N. Kittinger, A.M. Friedlander, S.D.  
884 Gaines, M.R. Clark, A.M. Thurnherr, S.M. Hardy and L.B. Crowder. 2015. Managing mining of  
885 the deep sea. *Science*. 349: 144-145.
- 886 2. Margolis, S.V. and R.G. Burns. 1976. Pacific deep-sea manganese nodules: Their distribution,  
887 composition, and origin. *Annual Review of Earth and Planetary Sciences*. 4: 229-263.
- 888 3. Hoagland, P., Beaulieu, S., Tivey, M.A., Eggert, E., German, C., Glowka, L., Link, J. 2010.  
889 Deep-sea mining of seafloor massive sulfides. *Marine Policy* 34: 728-32;  
890 [doi:10.1016/j.marpol.2009.12.001](https://doi.org/10.1016/j.marpol.2009.12.001)
- 891 4. Glasby, G.P. 2000. Lessons learned from deep-sea mining. *Science*. 289(5479): 551-553.
- 892 5. Burnett, W.C. and S.R. Riggs (Eds). 1987. *Phosphate Deposits of the World. Volume 3.*  
893 *Neogene to Modern Phosphorites*. Cambridge University Press, Cambridge, pp. 143-152.
- 894 6. Petersen, S., A. Krättschell and M.D. Hannington. 2016, March. The Current State of Global  
895 Activities Related to Deep-sea Mineral Exploration and Mining. In EAGE/DGG Workshop on  
896 Deep Mineral Exploration.
- 897 7. SPC. 2013. *Deep Sea Minerals: Deep Sea Minerals and the Green Economy*. Baker, E., and  
898 Beaudoin, Y. (Eds.) Vol. 2, Secretariat of the Pacific Community. Available at:  
899 [http://www.sopac.org/dsm/public/files/meetings/TrainingWorkshop4/UNEP\\_vol2.pdf](http://www.sopac.org/dsm/public/files/meetings/TrainingWorkshop4/UNEP_vol2.pdf)
- 900 8. Brake, L. and R. Peart. 2015. *Sustainable seas: managing the marine environment*. 400 p.  
901 Environmental Defence Society, New Zealand.
- 902 9. Mengerink, K.J., C.L. Van Dover, J. Ardron, M. Baker, E. Escobar-Briones, K. Gjerde, J.A.  
903 Koslow, E. Ramirez-Llodra, A. Lara-Lopez, D. Squires, T. Sutton, A.K. Sweetman and L.A.  
904 Levin. 2014. A Call for Deep-Ocean Stewardship. *Science*. 344: 696-698.
- 905 10. Jaeckel, A., J.A. Ardron and K.M. Gjerde. 2016. Sharing benefits of the common heritage of  
906 mankind – is the deep seabed mining regime ready? *Marine Policy*. 70: 198–204. doi:  
907 10.1016/j.marpol.2016.03.00.
- 908 11. Jaeckel, A. 2015. An Environmental Management Strategy for the International Seabed  
909 Authority? The Legal Basis. *The International Journal of Marine and Coastal Law*. 30: 93–119.
- 910 12. ISA. 2000. Regulations on Prospecting and Exploration for Polymetallic Nodules in the  
911 Area, ISBA/6/A/18, 13 July 2000.
- 912 13. ISA. 2010a. Regulations on Prospecting and Exploration for Polymetallic Sulphides in the  
913 Area (ISBA/16/A/12/Rev.1, 15 November 2010, as amended by ISBA/19/A/12, 25 July 2013  
914 and ISBA/20/A/10, 24 July 2014).
- 915 14. ISA. 2012. Regulations on Prospecting and Exploration for Cobalt-rich Ferromanganese  
916 Crusts in the Area (ISBA/18/A/11, 27 July 2012, as amended by ISBA/19/A/12, 25 July 2013).
- 917 15. ITLOS. 2011. Responsibilities and Obligations of States Sponsoring Persons and Entities with  
918 Respect to Activities in the Area (Seabed Disputes Chamber of the International Tribunal of  
919 the Law of the Sea, Case No 17, 1 February 2011).
- 920 16. Food and Agriculture Organization. 2009. *Management of Deep-sea Fisheries in the High*  
921 *Seas*. FAO, Rome, Italy.
- 922 17. Wedding, L.M., A.M. Friedlander, J.N. Kittinger, L. Watling, S.D. Gaines, M. Bennett, S.M.  
923 Hardy and C.R. Smith. 2013. From principles to practice: a spatial approach to systematic  
924 conservation planning in the deep sea. *Proceedings of the Royal Society of London B: .*  
925 *Biological Sciences*. 280(1773): 20131684. doi: 10.1098/rspb.2013.1684.
- 926 18. Ramirez-Llodra, E., P.A. Tyler, M.C. Baker, O.A. Bergstad, M.R. Clark, E. Escobar, L.A.  
927 Levin, L. Menot, A.A. Rowden, C.R. Smith and C.L. Van Dover. 2011. Man and the last great  
928 wilderness: Human impact on the deep sea. *PLOS One*. 6(8): e22588. doi:  
929 10.1371/journal.pone.0022588.
- 930 19. Levin, L.A. and N. Le Bris. 2015. Deep oceans under climate change. *Science*. 350: 766-768.
- 931 20. Groffman, P.M., J.S. Baron, T. Blett, A.J. Gold, I. Goodman, L.H. Gunderson, B.M. Levinson,  
932 M.A. Palmer, H.W. Paerl, G.D. Peterson and N.L. Poff. 2006. Ecological thresholds: the key to  
933 successful environmental management or an important concept with no practical application?  
934 *Ecosystems*. 9(1): 1-13.
- 935 21. Landres, P.B., P. Morgan and F.J. Swanson. 1999. Overview of the use of natural variability  
936 concepts in managing ecological systems. *Ecol. Applications* 9: 1179-1188.

- 937 22. Le, J., L.A. Levin and R.T. Carson. Incorporating ecosystem services into environmental  
938 management of deep-seabed mining. *Deep-sea Research II* (in press).
- 939 23. ISA. 2010b. A Geological Model of Polymetallic Nodule Deposits in the Clarion Clipperton  
940 Fracture Zone. International Seabed Authority Technical Report No. 6, Kingston, JA, 105 pp.  
941 ISBN: 978-976-95268-2-2
- 942 24. Smith, C.R., L.A. Levin, A. Koslow, P.A. Tyler and A.G. Glover. 2008c. The near future of the  
943 deep-sea floor ecosystems. In: Polunin, N.V.C. (Ed.) *Aquatic ecosystems: trends and global  
944 prospects*. Cambridge University Press, New York, 334-352.
- 945 25. Mullineaux, L.S. 1987. Organisms living on manganese nodules and crusts: distribution and  
946 abundance at three North Pacific sites. *Deep-Sea Research*. 34: 165-184.
- 947 26. Veillette, J., J. Sarrazin, A.J. Gooday, J. Galeron, J.-C. Caprais, A. Vangriesheim, J.  
948 Etoubleau, J.R. Christiand and S.K. Juniper. 2007. Ferromanganese nodule faunain the  
949 Tropical North Pacific Ocean: species richness, faunal cover and spatial distribution. *Deep-  
950 Sea Research I*. 54: 1912-1935.
- 951 27. Clark, M.; Smith, S. 2013. Environmental management considerations. Chapter 3. pp. 27-40.  
952 In: Baker, E., Bedouin, Y. (eds.). 1B. *Deep Sea Minerals: Manganese nodules-a physical,  
953 biological, environmental, and technical review*. SPC- GRIDArendal.
- 954 28. Smith, C.R., J. Galeron, A. Glover, A. Gooday, H. Kitazato, J. Lamshead, L. Menot, G.  
955 Paterson, A. Rogers and M. Sibuet. 2008b. Biodiversity, species ranges, and gene flow in the  
956 abyssal Pacific nodule province: predicting and managing the impacts of deep seabed mining.  
957 ISA Technical Study no. 3, International Seabed Authority, Kingston, Jamaica.
- 958 29. Vanreusel, A., A. Hilario, P.A. Ribeiro, L. Menot and P. Martinez Arbizu. 2016. Threatened by  
959 mining, polymetallic nodules are required to preserve abyssal epifauna. *Scientific Reports*. 6.  
960 doi: 10.1038/srep26808.
- 961 30. Amon, D.J., A.F. Ziegler, T. Dahlgren, A. Glover, A. Goineau, A.J. Gooday, H. Wiklund and  
962 C.R. Smith. 2016. Insights into the abundance and diversity of abyssal megafauna in a  
963 polymetallic-nodule region in the eastern Clarion-Clipperton Zone. *Nature Scientific Reports*, in  
964 press.
- 965 31. Khripounoff, A., J.C. Caprais, P. Crassous and J. Etoubleau. 2006. Geochemical and  
966 biological recovery of the disturbed seafloor in polymetallic nodule fields of the Clipperton-  
967 Clarion Fracture Zone (CCFZ) at 5000 m depth. *Limnology and Oceanography*. 51 (5): 2033-  
968 2041.
- 969 32. Berelson, W.M., R.F. Anderson, J. Dymond, D. DeMaster, D.E. Hammond, R. Collier, S.  
970 Honjo, M. Leinen, J. McManus, R. Pope and C. Smith. 1997. Biogenic budgets of particle rain,  
971 benthic remineralization and sediment accumulation in the equatorial Pacific. *Deep-Sea  
972 Research*. 44: 2251-22821. doi: 10.1016/S0967-0645(97)00030-1.
- 973 33. Smith, C.R. and A.W.J. Demopoulos. 2003. Ecology of the deep Pacific Ocean floor. In: Tyler,  
974 P.A. (Ed.) *Ecosystems of the World, Ecosystems of the Deep Ocean*, vol. 28. Elsevier,  
975 Amsterdam, pp. 179 – 218.
- 976 34. Yamazaki, T., Y. Kajitani, B. Barnett, T. and Suzuki. 1997. "Development of Image Analytical  
977 Technique for Resedimentation Induced by Nodule Mining," *Proc 2nd ISOPE Ocean Mining  
978 Symp*, Seoul, Korea, International Society of Offshore and Polar Engineers, pp. 159-164.
- 979 35. Cronan, D.S. 1999. Review of geochemical impacts of polymetallic nodule mining. In: (Eds.  
980 ISA) *Deep-Seabed Polymetallic Nodule Exploration: Development of Environmental  
981 Guidelines*. International Seabed Authority, Kingston, Jamaica. 118-154.
- 982 36. Gardner, W.D., L.G. Lawrence, and E. M. Thorndike. 1984. Long-term photographic, current  
983 and nephelometer observations of manganese nodule environments in the Pacific. *Earth and  
984 Planetary Sci. Lett*. 70: 95-109.
- 985 37. Southall, B.L. 2005. Final report of the NOAA International Symposium on "Shipping noise  
986 and marine mammals: a forum for science, management, and technology. May 2004".  
987 [http://www.nmfs.noaa.gov/pr/pdfs/acoustics/shipping\\_noise.pdf](http://www.nmfs.noaa.gov/pr/pdfs/acoustics/shipping_noise.pdf)
- 988 38. Van Dover, C.L. 2014. Impacts of anthropogenic disturbances at deep-sea hydrothermal vent  
989 ecosystems: A review. *Marine Environmental Research*. 102: 59-72.

- 990 39. Smith, C.R., F.C. De Leo, A.F. Bernardino, A.K. Sweetman and P. Martinez Arbizu. 2008a.  
991 Abyssal food limitation, ecosystem structure and climate change. *Trends in Ecology and*  
992 *Evolution*. 23: 518-528.
- 993 40. Hein, J. and S. Petersen. 2013. The geology of cobalt-rich ferromanganese crusts. In: Baker,  
994 E., Beudoin, Y. (ed) *Deep Sea Minerals: Cobalt-rich ferromanganese crusts; a physical,*  
995 *biological, environmental, and technical review*. vol 1C. Secretariat of the Pacific Community,  
996 pp. 7-14.
- 997 41. Miljutin, D.M., M. Miljutina, P. Martinez Arbizu and J. Galeron. 2011. Deep-sea nematode  
998 assemblage has not recovered 26 years after experimental mining of polymetallic nodules  
999 (Clarion-Clipperton Fracture Zone, Tropical Eastern Pacific). *Deep-Sea Research I*. 58: 885-  
1000 897.
- 1001 42. Hannington, M., J. Jamieson, T Monecke, S. Petersen and S. Beaulieu. 2011. The abundance  
1002 of seafloor massive sulfide deposits. *Geology*. 39: 1155-1158. Hoagland, P., S. Beaulieu, M.A.  
1003 Tivey, R.G. Eggert, C. German, L. Glowka, L. and J. Lin. 2010. Deep-sea mining of seafloor  
1004 massive sulfides. *Marine Policy*. 34(3): 728-732.
- 1005 43. Boschen, R.E., A.A. Rowden, M.R. Clark and J.P.A. Gardner. 2013. Mining of deep-sea  
1006 seafloor massive sulfides: A review of the deposits, their benthic communities, impacts from  
1007 mining, regulatory frameworks and management strategies. *Ocean and Coastal Management*.  
1008 84: 54-67.
- 1009 44. Van Dover, C.L., Grassle, J.F. and M. Boudrias. 1990. Hydrothermal vent fauna of Escanaba  
1010 trough (Gorda Ridge). In: McMurray, G.R (Ed.) *Gorda Ridge: A seafloor spreading center in*  
1011 *the United States' exclusive economic zone*. Springer, New York, pp. 285-287.
- 1012 45. Erickson, K.L., S.A. Macko and C.L. Van Dover. 2009. Evidence for a chemoautotrophically  
1013 based food web at inactive hydrothermal vents (Manus Basin). *Deep-Sea Research II*. 56:  
1014 1577-1585.
- 1015 46. Boschen, R.E., Rowden, A.A., Clark, M.R., Pallentin, A., Gardner, J.P.A. (2016) Seafloor  
1016 massive sulfide deposits support unique megafauna assemblages: implications for seabed  
1017 mining and conservation. *Marine Environmental Research* 115: 78-88.
- 1018 47. Beaulieu, S.E., E.T. Baker and C.R. German. 2015. Where are the undiscovered  
1019 hydrothermal vents on oceanic spreading ridges? *Deep Sea Research Part II: Topical Studies*  
1020 *in Oceanography*. 121: 202-212.
- 1021 48. Baker, E., J. Resing, R. Haymon, V. Tunnicliffe, R. Martinez, V. Ferrini, S. Walker and K.  
1022 Nakamura. 2016. How many vent fields? New estimates of vent field populations on ocean  
1023 ridges from precise mapping of hydrothermal discharge locations. *Earth and Planetary*  
1024 *Science Letters*. 449: 186-196.
- 1025 49. Levin, L.A., G.F. Mendoza, T. Konotchick and R. Lee. 2009. Community structure and trophic  
1026 relationships in Pacific hydrothermal sediments. *Deep-Sea Research II*. 56: 1632-1648
- 1027 50. Van Dover, C.L. 2011. Tighten regulations on deep- sea mining. *Nature*. 470(7332): 31-33.  
1028 doi: 10.1038/470031a.
- 1029 51. Smith, S. and R. Heydon. 2013. Processes related to the technical development of marine  
1030 mining. In: Baker, E., Beudoin, Y. (eds.) *Deep Sea Minerals: Cobalt-rich ferromanganese*  
1031 *crusts; a physical, biological, environmental, and technical review*. vol 1C. Secretariat of the  
1032 *South Pacific Community*, 41-46.
- 1033 52. Van Dover, C.L. 2010. Mining seafloor massive sulphides and biodiversity: what is at risk?  
1034 *ICES Journal of Marine Science*. 68(2): 341-348. doi: 10.1093/icesjms/fsq086.
- 1035 53. Coffey Natural Systems. 2008. Environmental Impact Statement. Solwara 1 project Nautilus  
1036 Minerals Niugini Limited, Main Report. Brisbane: Coffey Natural Systems. Available at:  
1037 [http://www.nautilusminerals.com/irm/content/pdf/environmental-](http://www.nautilusminerals.com/irm/content/pdf/environmental-reports/Environmental%20Impact%20Statement%20-%20Main%20Report.pdf)  
1038 [reports/Environmental%20Impact%20Statement%20-%20Main%20Report.pdf](http://www.nautilusminerals.com/irm/content/pdf/environmental-reports/Environmental%20Impact%20Statement%20-%20Main%20Report.pdf)
- 1039 54. Tunnicliffe, V., R. W. Embley, J. F. Holden, D.A. Butterfield, G.J. Massoth and S.K. Juniper.  
1040 1997. Biological colonization of new hydrothermal vents following an eruption on Juan de Fuca  
1041 Ridge. *Deep Sea Research I*. 44: 1627-1644.
- 1042 55. Nakajima, R., H. Yamamoto, S. Kawagucci, Y. Takaya, T. Nozaki, C. Chen, K. Fujikura, T.  
1043 Miwa and K. Takai. 2015. Post-drilling changes in seabed landscape and megabenthos in a



- 1044 deep-sea hydrothermal system, the Iheya North field, Okinawa Trough. *PLoS one*. 10(4):  
1045 e0123095.
- 1046 56. Van Dover, C.L. 2000. *The Ecology of Deep-Sea Hydrothermal Vents*. Princeton University  
1047 Press. Princeton, New Jersey, USA. 448 pp.
- 1048 57. Baker, M.C., E. Ramirez-Llodra, P.A. Tyler, C.R. German, A. Boetius, E.E. Cordes, N.  
1049 Dubilier, C.R. Fisher, L. Levin, A. Metaxas, A. Rowden, R.S. Santos, T. Shank, C. Van Dover,  
1050 C. Young and A. Waren. 2010. Biogeography, Ecology and Vulnerability of Chemosynthetic  
1051 Ecosystems in the Deep Sea. Chapter 9. In: McIntyre, A.D. (Ed.) *Life in the World's Oceans*.  
1052 Blackwell Publishing Ltd., pp. 161–182.
- 1053 58. Ramirez-Llodra, E., A. Brandt, R. Danovaro, B. De Mol, E. Escobar, C.R. German, L.A. Levin,  
1054 P.M. Arbizu, L. Menot, P. Buhl-Mortensen, B.E. Narayanaswamy, C.R. Smith, D.P. Tittensor,  
1055 P.A. Tyler, A. Vanreusel and M. Vecchione. 2010. Deep, diverse and definitely different:  
1056 unique attributes of the world's largest ecosystem. *Biogeosciences*. 7(9): 2851-2899. doi:  
1057 10.5194/bg-7-2851-2010.
- 1058 59. Levin, L.A., A.R. Baco, D.A. Bowden, A. Colaco, E.E. Cordes, M.R. Cunha, A.W.J.  
1059 Demopoulos, J. Gobin, B.M. Grupe, J. Le and A. Metaxas. 2016. Hydrothermal Vents and  
1060 Methane Seeps: Rethinking the Sphere of Influence. *Frontiers of Marine Science*. 3: 72.
- 1061 60. Hein, J.R. 2004. Cobalt-rich ferromanganese crusts: Global distribution, composition, origin  
1062 and research activities. In: *Minerals other than polymetallic nodules of the International  
1063 Seabed Area; Proceedings of a workshop held on 26-30 June 2000, International Seabed  
1064 Authority, Kingston, Jamaica, volume 1, 188-256.*
- 1065 61. Yesson, C., M.R. Clark, M.L. Taylor and A.D. Rogers. 2011. The global distribution of  
1066 seamounts based on 30 arc seconds bathymetry data. *Deep-Sea Research Part I-  
1067 Oceanographic Research Papers*. 58(4): 442-453. doi: 10.1016/j.dsr.2011.02.004.
- 1068 62. Hein, J.R., A. Koschinsky, M. Bau, F.T. Manheim, J.-K. Kang and L. Roberts. 2000. Cobalt-  
1069 rich ferromanganese crusts in the Pacific. In: Cronan, D. S. (ed) *Handbook of marine mineral  
1070 deposits*. CRC Press, Boca Raton, Florida, 239-279.
- 1071 63. Hein, J.R., T.A. Conrad and H. Staudigel. 2010. Seamount Mineral Deposits a Source of Rare  
1072 Metals for High-Technology Industries. *Oceanography* 23(1):184-189.
- 1073 64. Hein, J.R., T.A. Conrad and R.E. Dunham. 2009. Seamount characteristics and mine-site  
1074 model applied to exploration and mining lease block selection for cobalt-rich ferromanganese  
1075 crusts. *Mar. Georesour. Geotec.* 27(2): 160-176. doi: 10.1080/10641190902852485.
- 1076 65. He, G., W. Ma, C. Song, S. Yang, B. Zhu, H. Yao, X. Jiang and Y. Cheng. 2011. Distribution  
1077 characteristics of seamount cobalt-rich ferromanganese crusts and the determination of the  
1078 size of areas for exploration and exploitation. *Acta Oceanol. Sin.* 30(3): 63-75.
- 1079 66. Clark, M.R. 2013. Biology associated with cobalt-rich ferromanganese crusts. In: Baker, E.,  
1080 Beudoin, Y. (ed) *Deep Sea Minerals: cobalt-rich ferromanganese crusts; a physical, biological,  
1081 environmental, and technical review*. vol 1C. Secretariat of the Pacific Community, 15-22.
- 1082 67. Levin, L.A. and C.L. Thomas. 1988. The ecology of xenophyophores (Protista) on eastern  
1083 Pacific seamounts. *Deep-Sea Research*. 35(12): 2003-2027.
- 1084 68. Williams, A., T.A. Schlacher,  
1085 A.A. Rowden, F. Althaus, M.R. Clark, D.A. Bowden, R. Stewart, N.J. Bax, M. Consalvey and  
1086 R.J. Kloser. 2010. Seamount megabenthic assemblages fail to recover from trawling impacts.  
1087 *Marine Ecology*. 31: 183-199.
- 1088 68. Clark, M.R., A.A. Rowden, T. Schlacher, A. Williams, M. Consalvey, K.I. Stocks, A.D. Rogers,  
1089 T.D. O'Hara, M. White, T.M. Shank and J.M. Hall-Spencer. 2010. The ecology of seamounts:  
1090 Structure, function, and human impacts. *Annual Review of Marine Science*. 2:253-278.
- 1091 69. Buhl-Mortensen, L., A. Vanreusel, A.J. Gooday, L.A. Levin, I.G. Priede, P. Buhl-Mortensen, H.  
1092 Gheerardyn, N.J. King and M. Raes. 2010. Biological structures as a source of habitat  
1093 heterogeneity and biodiversity on the deep ocean margins. *Marine Ecology*. 31: 21-50.
- 1094 70. Williams, A., T. A. Schlacher, A. A. Rowden, F. Althaus, M. R. Clark, D. A. Bowden, R.  
1095 Stewart, N. J. Bax, M. Consalvey and R. J. Kloser. 2010. Seamount megabenthic assemblages  
1096 fail to recover from trawling impacts. *Marine Ecology* 31:183-199.
- 1097 71. Clark, M.R., F. Althaus, T.A. Schlacher, A. Williams, D.A. Bowden and A.A. Rowden. 2015.  
1098 The impacts of deep-sea fisheries on benthic communities: a review. *ICES Journal of Marine  
1099 Science*. 73: i51-i69. doi: 10.1093/icesjms/fsv123. Clark, M. and S. Smith. 2013. Environmental  
management considerations. Chapter 3. pp. 27-42. In: Baker, E., Bedouin, Y. (eds.). 1A. Deep

- 1100 Sea Minerals: Sea-floor massive sulphides-a physical, biological, environmental, and technical  
1101 review. Secretariat of the Pacific Community.
- 1102 72. Yamazaki, T., E. Kuboki and T. Matsui. 2001. DIETS: a new benthic impact experiment on a  
1103 seamount. In: Fourth Ocean Mining Symposium. Szczecin, Poland, pp. 69-76.
- 1104 73. IUCN. 2015. Deep-sea mining: environmental issues associated with deep-sea minerals  
1105 exploration. IUCN, NIWA, TIS.
- 1106 74. Schlacher, T.A., A.R. Baco, A.A. Rowden, T.D. O'Hara, M.R. Clark, C. Kelley and J.F. Dower.  
1107 2014. Seamount benthos in a cobalt-rich crust region of the central Pacific: Conservation  
1108 challenges for future seabed mining. *Diversity and Distributions*. 20(5): 491-502.
- 1109 75. Morgan, N.B., S. Cairns, H. Reiswig and A.R. Baco. 2015. Benthic mega-faunal community  
1110 structure of cobalt-rich manganese crusts on Necker Ridge. *Deep-Sea Research I*. 104: 92-  
1111 105.
- 1112 76. Rogers, A.D., A. Baco, H. Griffiths, T. Hart and J.M. Hall-Spencer. 2007. Corals on  
1113 Seamounts. In: Pitcher, T.J., T. Morato, P.J.B. Hart, M.R. Clark, N. Haggan and R. S. Santos  
1114 (eds) Seamounts: Ecology, Fisheries and Conservation. Blackwell Publishing Ltd., pp. 141-  
1115 169.
- 1116 77. Rowden, A.A., T.A. Schlacher, A. Williams, M.R. Clark, R. Stewart, F. Althaus, D.A. Bowden,  
1117 M. Consalvey, W. Robinson and J. Dowdney. 2010b. A test of the seamount oasis hypothesis:  
1118 seamounts support higher epibenthic megafaunal biomass than adjacent slopes. *Marine  
1119 Ecology- An Evolutionary Perspective*. 31: 95-106. doi: 10.1111/j.1439-0485.2010.00369.x.
- 1120 78. Freiwald, A., J.H. Fossa, A.J. Grehan, J.A. Koslow and J.M. Roberts. 2004. Cold-water coral  
1121 reefs: out of sight - no longer out of mind. UNEP World Conservation Monitoring Centre.  
1122 UNEP-WCMC, Cambridge, UK, 84 pp.
- 1123 79. Costello, M.J., M. McCrea, A. Freiwald, T. Lundalv, L. Jonsson, B.J. Bett, T.C.E. van Weering,  
1124 H. de Haas, J.M. Roberts and D. Allen. 2005. Role of cold-water *Lophelia pertusa* coral reefs  
1125 as fish habitat in the NE Atlantic. In: Freiwald, A. and J.M. Roberts (eds) Cold-water corals and  
1126 ecosystems. Springer-Verlag, Germany, pp. 771-805.
- 1127 80. Althaus, F., A. Williams, T.A. Schlacher, R.K. Kloser, M.A. Green, B.A. Barker, N.J. Bax, P.  
1128 Brodie and M.A. Schlacher-Hoenlinger. 2009. Impacts of bottom trawling on deep-coral  
1129 ecosystems of seamounts are long-lasting. *Marine Ecology Progress Series*. 397: 279-294.
- 1130 81. Clark, M.R., C. Kelley, A. Baco and A.A. Rowden. 2011. Fauna of cobalt-rich ferromanganese  
1131 crust seamounts. International Seabed Authority Technical Study No. 8. International Seabed  
1132 Authority (ISA), Kingston, Jamaica.
- 1133 82. Genin, A. and J.F. Dower. 2007. Seamount plankton dynamics. In: Pitcher, T.J., T. Morato,  
1134 P.J.B. Hart, M.R. Clark, N. Haggan and R.S. Santos (eds) Seamount: Ecology, Fisheries and  
1135 Conservation. Blackwell, Oxford, pp. 86-100.
- 1136 83. Genin A., P.K. Dayton, P.F. Lonsdale, and F.N. Spiess. 1986. Corals on seamount peaks  
1137 provide evidence of current acceleration over deep-sea topography. *Nature* 322:59-61.
- 1138 84. Morato T., C. Bulman, and T.J. Pitcher. 2009. Modelled effects of primary and secondary  
1139 production enhancement by seamounts on local fish stocks. *Deep-Sea Research Part II*  
1140 56:2713-2719.
- 1141 85. Rowden, A.A., J.F. Dower, T.A. Schlacher, M. Consalvey and M.R. Clark. 2010a. Paradigms  
1142 in seamount ecology: fact, fiction and future. *Marine Ecology- An Evolutionary Perspective*. 31:  
1143 226-241. doi: 10.1111/j.1439-0485.2010.00400.x.
- 1144 86. Stocks, K. and P.J.B. Hart. 2007. Biogeography and biodiversity of seamounts. In: Pitcher,  
1145 T.J., T. Morato, P.J.B. Hart, M.R. Clark, N. Haggan and R.S. Santos (eds.) Seamounts:  
1146 Ecology, Fisheries and Conservation. Blackwell, Oxford, pp. 255-281.
- 1147 87. Baturin, G.N. 1982. Phosphorites on the seafloor. Origin, composition and distribution.  
1148 *Developments in Sedimentology*. 33: 1-343.
- 1149 88. Föllmi, K.B. 1996. The phosphorus cycle, phosphogenesis and marine phosphate-rich  
1150 deposits. *Earth-Science Reviews*. 40: 55-124.
- 1151 89. McKelvey, V.E. 1967. Phosphate deposits. USGS Report Series 1252-D. U.S. Government  
1152 Printing Office. 21 pp.
- 1153 90. Thiel, H., M.V. Angel, E.J. Foell, A.L. Rice, G. Schriever and A. Below. 1998. Marine science  
1154 and technology – Environmental risks from large-scale ecological research in the deep sea: a

- 1155 desk study. Section 2.5: pp 67-78. Luxembourg: Office for Official Publications of the  
1156 European Communities. ISBN 92-828-3517-0.
- 1157 91. Schulz, H.N. and H.D. Schulz. 2005. Large sulfur bacteria and the formation of  
1158 phosphorite. *Science*. 307(5708): 416-418.
- 1159 92. Goldhammer, T., V. Brüchert, T.G. Ferdelman and M. Zabel. 2010. Microbial sequestration of  
1160 phosphorus in anoxic upwelling sediments. *Nature Geoscience*. 3(8): 557-561. doi:  
1161 10.1038/NGEO913.
- 1162 93. Brock, J. and H.N. Schulz-Vogt. 2011. Sulfide induces phosphate release from polyphosphate  
1163 in cultures of a marine *Beggiatoa* strain. *The ISME Journal*. 5(3): 497-506.
- 1164 94. Crosby, C.H. and J.V. Bailey. 2012. The role of microbes in the formation of modern and  
1165 ancient phosphatic mineral deposits. *Frontiers in Microbiology*. 3(241): 1-6. doi:  
1166 10.3389/fmicb.2012.00241
- 1167 95. Bremner, J.M. and J. Rogers. 1990. Phosphorite deposits on the Namibian continental shelf.  
1168 In: *Phosphate Deposits of the World Vol. 3* (eds Burnett, W. C. & Riggs, S. R.) 143–152  
1169 Cambridge Univ. Press.
- 1170 96. Coles, S.K.P., C.I. Wright, D.A. Sinclair and P.V.D Bossche. 2002. The Potential for  
1171 Environmentally Sound Development of Marine Deposits of Potassic and Phosphatic Minerals  
1172 Offshore Southern Africa. *Marine Georesources and Geotechnology*. 20(2): 87-110. doi:  
1173 10.1080/03608860290051822.
- 1174 97. Dawson, E. 1984. The benthic fauna of the Chatham Rise: An assessment relative to possible  
1175 effects of phosphorite mining. *Geologisches Jahrbuch*. D65: 209–231.
- 1176 98. Gnandi, K., G. Tchagnbedji, K. Killi, G. Baba and K. Abbe. 2006. The impact of phosphate  
1177 mine tailings on the bioaccumulation of heavy metals in marine fish and crustaceans from the  
1178 coastal zone of Togo. *Mine Water and the Environment*. 25: 56–62
- 1179 99. Leduc, D., A.A. Rowden, L.G. Torres, S.D. Nodder and A. Pallentin. 2015. Distribution of  
1180 macro-infaunal communities in phosphorite nodule deposits on Chatham Rise, Southwest  
1181 Pacific: Implications for management of seabed mining. *Deep-Sea Research I*. 99: 105–118.
- 1182 100. Utne-Palm A.C., A.G.V. Salvanes, B. Currie, S. Kaartvedt, G.E. Nilsson, V.A. Braithwaite,  
1183 J.A.W. Stecyk, M. Hundt, M. van der Bank, B. Flynn, G.K. Sandvik, T.A. Klevjer, A.K.  
1184 Sweetman, V. Brüchert, K. Pittman, K.R. Peard, I.G. Lunde, R.A.U. Strandabø and M.J.  
1185 Gibbons. 2010. Trophic Structure and Community Stability in an Overfished Ecosystem.  
1186 *Science*. 329: 333. doi: 10.1126/science.1190708.
- 1187 101. Halfar J. and R.M. Fujita. 2002. Precautionary management of deep-sea mining. *Marine*  
1188 *Policy*. 26: 103–106.
- 1189 102. Van Dover, C.L, C.R. Smith and L. Godet. 2011. Chemosynthetic Ecosystems:  
1190 Understanding What's at Risk and Tools for Effective Management. *Marine Protected Areas at*  
1191 *the the High Seas - Symposium.*, Feb. 2011, Londres, United Kingdom.
- 1192 103. Helson, J., S. Leslie, G. Clement, R. Wells and R. Wood. 2010. Private rights, public  
1193 benefits: industry-driven seabed protection. *Marine Policy*. 34(3): 557-566. ISA. 2000.  
1194 Regulations on Prospecting and Exploration for Polymetallic Nodules in the Area  
1195 (ISBA/6/A/18, 13 July 2000, as amended by ISBA/19/A/9 and ISBA/19/A/12, 25 July 2013, and  
1196 ISBA/20/A/9, 24 July 2014).
- 1197 104. Ramirez-Llodra, E., H.C. Trannum, M. Schaanning, A. Evenset, B. Flem, T.E. Finne, M.  
1198 Andersson, L.A. Levin, A. Vanreusel and A. Hilario. 2015. Submarine and deep-sea mine  
1199 tailing placements: a review of current practices, environmental issues, natural analogs and  
1200 knowledge gaps in Norway and internationally. *Marine Pollution Bulletin*. 97: 13-35.  
1201 <http://dx.doi.org/10.1016/j.marpolbul.2015.05.062>
- 1202 105. Bopp, L., L. Resplandy, J.C. Orr, S.C. Doney, J.P. Dunne, M. Gehlen, P. Halloran, C.  
1203 Heinze, T. Ilyina, R. Seferian and J. Tjiputra. 2013. Multiple stressors of ocean ecosystems in  
1204 the 21st century: projections with CMIP5 models. *Biogeosciences*. 10: 6225–6245. doi:  
1205 10.5194/bg-10-6225-2013.
- 1206 106. Mora, C., C.-L. Wei, A. Rollo, T. Amaro, A.R. Baco, D. Billett, L. Bopp, Q. Chen, M. Collier,  
1207 R. Danovaro, A.J. Gooday, B.M. Grube, P.R. Halloran, J. Ingels, D.O.B. Jones, L.A. Levin, H.  
1208 Nakano, K. Norling, E. Ramirez-Llodra, M. Rex, H.A. Ruhl, C.R. Smith, A.K. Sweetman, A.R.  
1209 Thurber, J.F. Tjiputra, P. Usseglio, L. Watling, T. Wu and M. Yasuhura. 2013. Biotic and

1210 human vulnerability to projected changes in ocean biogeochemistry over the 21<sup>st</sup> Century.  
1211 PLoS Biology 11(10): e1001682. doi: 10.1371/journal.pbio.1001682.  
1212 107. ISA and ITLOS Preparatory Commission. 1990. Draft Regulations on Prospecting,  
1213 Exploration and Exploitation of Polymetallic Nodules in the Area.  
1214 LOS/PCN/SCN.3/WP.6/Add.5 (8 February 1990), article 2(2).  
1215 108. Synnes, M. 2007. Bioprospecting of organisms from the deep sea: scientific and  
1216 environmental aspects. Clean Technologies and Environmental Policy. 9(1): 53-59.  
1217 109. Yao, H., M. Dao, T. Imholt, J. Huang, K. Wheeler, A. Bonilla, S. Suresh and C. Ortiz. 2010.  
1218 Protection mechanisms of the iron-plated armor of a deep-sea hydrothermal vent gastropod.  
1219 Proc. Natl. Acad. Sci 107: 987-992. doi: 10.1073/pnas.0912988107.  
1220 110. Mahon, B.P., A. Bhatt, D. Vulla, C.T. Supuran and R. McKenna. 2015. Exploration of an  
1221 ionic inhibition of the  $\alpha$ -carbonic anhydrase from *Thiomicrospiracrunogena* XCL-2  
1222 gammaproteobacterium: a potential bio-catalytic agent for industrial CO<sub>2</sub> removal. Chem. Eng.  
1223 Sci. 138: 575–580. doi: 10.1016/j.ces.2015.07.030  
1224 111. Lee, S.Y. and G.D. Holder. 2001. Methane hydrates potential as a future energy  
1225 source. Fuel Processing Technology. 71(1): 181-186.  
1226  
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1228 **Figure Captions.**

1229 **Figure 1. Areas beyond national jurisdiction that have been claimed for minerals mining exploration.**

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